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Final Report
for
DEVELOPMENT OF DATA BASE FOR
HUMAN EXPOSURE TO AIR POLLUTION
IN THE SOUTH COAST AIR BASIN

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1. INTRODUCTION AND SUMMARY

The purpose of this project is to develop a common data base of human exposure to air pollution in the South Coast Air Basin (SCAB), so that the same data base can be used for various epidemiological studies of air pollution health effects. In the past studies, the level of each individual's exposure to air pollution was assumed to be the same as the exposure level of the community from which a controlled (i.e., study) population was selected. However, as an air pollution level varies with time and space, so does an exposure level of a person, depending upon his spatial position, daily activity pattern, and other factors.

Some health effects are considered to be caused more by a short-term peak (acute) exposure rather than by a long-term average (chronic) exposure. The opposite may be true for some other health effects. Therefore, to perform a comprehensive epidemiological study requires both short-term and long-term exposure estimates. People are exposed simultaneously to several pollutants. Simultaneous exposures to multiple pollutants may cause either synergistic (more than additive) or antagonistic (less than additive) effects on human health. Therefore, the developed data base contains human exposure data for each of the six major (four gaseous and two particulate) criteria pollutants. To help investigate those health effects which have a long latency period, the data base covers an extended period of eleven years, from January 1966 through December 1976.

1.1 Outline of Data Base

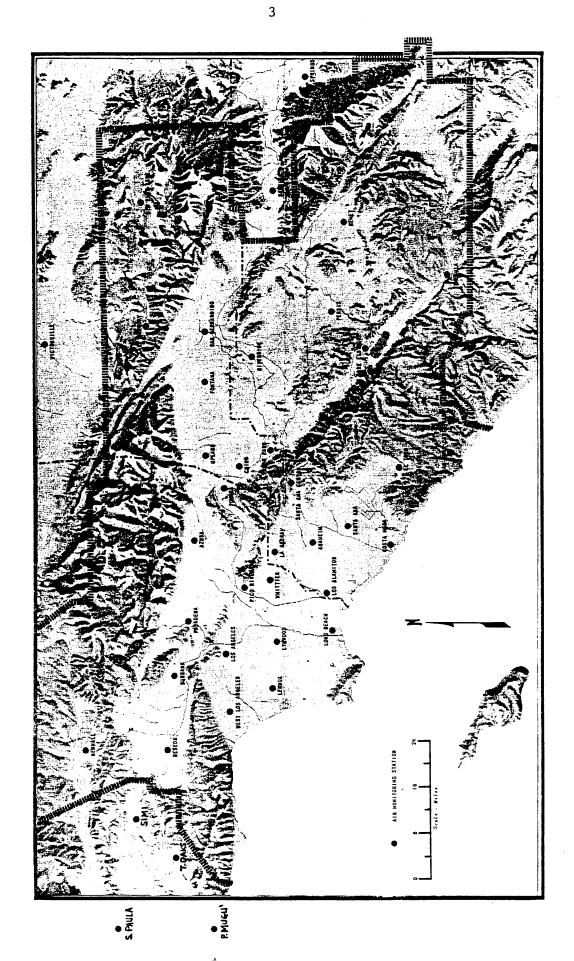
A large human exposure data base has been developed to provide a convenient human exposure data source for performing an epidemiological study of air pollution health effects. The data base includes monthly values of several key exposure parameters, computed for each of the four gaseous (OX, NO2, SO2 and CO) and two particulate (TSP and SO4)

pollutants at some 400 postal zip-code areas as well as at some 40 air monitoring stations scattered over the SCAB (see Figure 1 and Attachment F). Hourly concentration data were used to compute exposure values for the gaseous pollutants, while 24-hour concentration data were used for the particulate pollutants.

Given the residence location of a person by zip-code, one can readily extract from the data base necessary information about outdoor exposure characteristics of that person for any month or period during the 11-year period. Given UTM coordinates of a person or a group of people, one can easily compute from the data base monthly exposure values by applying the attached spatial interpolation program to the human exposure data at air monitoring sites. Given a history of a person's residences and/or work places, one can compute from the data base both cumulative exposure values and a history of exposures by month.

A certain adverse health effect may be caused either by a short-term exposure to peak concentrations or by a long-term exposure to low concentrations. To cope with such uncertainty in the air pollution-health effects relationship, the exposure data are computed for both total exposures and excess exposures. Here, an excess exposure means that a person is exposed to pollutant concentrations above a given concentration threshold. For each pollutant, four different concentration thresholds were chosen by considering the known (but not quantified) health effects from exposures to that pollutant.

The total exposure is given each month by a mean concentration and an apparent dose (i.e., time integral of concentration). The excess exposure is given by an exceedance frequency (i.e., number of hours exceeding a given threshold) and an excess dose (i.e., time integral of concentrations exceeding a threshold level). Both the total and excess exposure data are computed each month for each pollutant at both the monitor site and the centroid of each postal zip-code area. Exposure data at each zip-code area are labeled by a confidence indicator, which was determined by computing the distance between the centroid of that zip-code area and the nearest monitoring station with valid air quality data. If the distance is less than a



Location of Air Monitoring Stations Used for Developing Human Exposure Data Bases in SCAB Region Figure 1.

IAI

typical representative area of a monitor for a given pollutant, the confidence indicator for that zip-code area and pollutant is denoted as A, the highest confidence level. If the distance is greater than the typical representative area but less than twice the typical representative area, the confidence indicator, B, is used to indicate the moderate confidence in the data. If the distance is greater than twice the typical representative area, C is used to indicate that the exposure data have the least confidence.

1.2 Recommended Use of the Data Base

The developed data base provides spatial and temporal characteristics of <u>apparent</u> exposure, i.e., that estimated from ambient air quality monitoring data, instead of <u>true</u> exposure to air pollution. Therefore, if one suspects that exposure to a certain pollutant occurs mostly indoors rather than outdoors, the data base should not be used for such a case. However, if human exposure occurs partially indoors and partially outdoors, the data base can be used to approximate the outdoor exposure. In most epidemiological studies, there is no praticable means of measuring the indoor exposure. It is recommended that a substitutable quantity be found from information obtainable from survey questionnaires about daily activity patterns and housing characteristics.

Given a person's residence history, his cumulative exposure can be computed easily by retrieving from the data base a set of exposure data for his residence in Period Pl and Zip-Code Area Zl, another set of exposure data in Period P2 and Zip-Code Area Z2, and so on. All we have to do is just sum those exposure values over the entire study period. If one wants to know seasonal exposure, he can add the monthly exposure values for those months that belong to the same season.

If one would like to refine the exposure estimate for a person whose residence location and work place are both known, the following procedure is recommended: First, all day exposures at the person's residence and work place are computed from the data base. Second, the

exposure value at his residence is adjusted for weekend/weekday air quality differences by multiplying it by 0.93 for photochemical oxidants, 1.0 for sulfate, and 0.75 for the other four pollutants [Horie et al. 1979]. Then, the person's exposure is estimated by taking a weighted average of the two exposure values, i.e.,

[5(exposure at work place) + 2(exposure at residence)] ÷ 7

The above procedure of refining an exposure estimate will work rather well for photochemical oxidants, whose nightime concentrations are virtually zero. For other pollutants whose nightime levels can be high, the above procedure will not work as well as for photochemical oxidants. The reason is that the recommended procedure does not take into account a difference between nighttime exposure at the residence location and that at the work place. To correct the defect of the recommended procedure may require a further refinement of indoor/outdoor exposures, which is beyond the scope of the present study.

In some cases, one would like to have exposure parameter values for a threshold that is different from those used in the data base. In such cases, one can use either the percentile concentration data tabulated for the gaseous pollutants or the original concentration data tabulated for the particulate pollutants. By applying the following human-exposure program to the percentile or original concentration data, one can compute exposure parameter values for a desired concentration threshold. In some other cases, one may want to have exposure data at regularly spaced grid points instead of at the centroid of each zip-code area. In such cases, one can use the exposure data computed at each air monitoring station and the spatial interpolation computer program attached to this data base. All he has to do is to apply the spatial interpolation program to the station exposure data and compute the desired exposure data at each grid point.

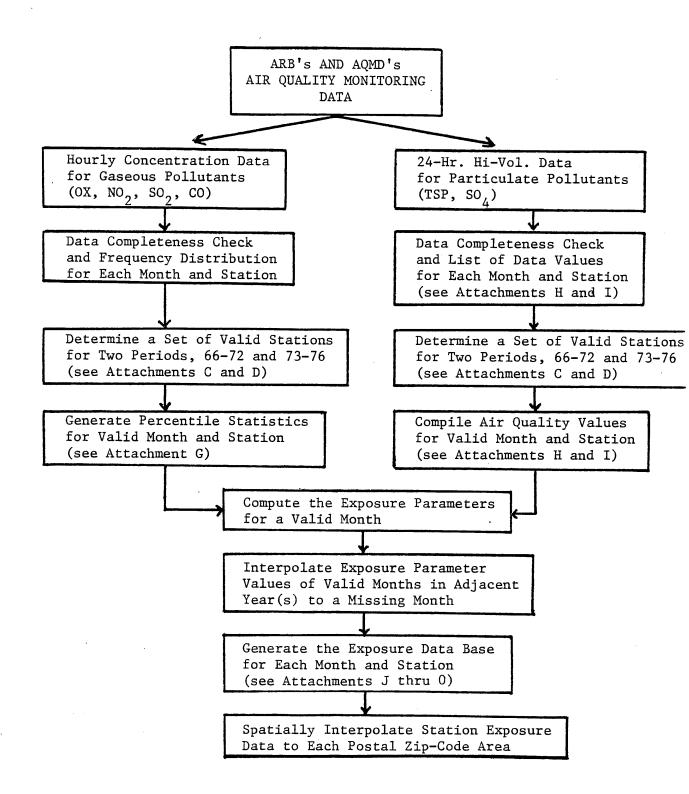


Figure 2. Flow Diagram of Data Processing Steps from Original
Air Quality Data to Human Exposure Data (see Attachment A)

Table 1. Number of Valid Stations* Used for Developing
Human Exposure Data during the 1st (1966-72)
and 2nd (1973-76) Periods.

	·· ·· · · · · · · · · · · · · · · ·	
Pollutant	1966-1972	1973-1976
Photochemical Oxidants (OX)	11	30
Nitrogen Dioxide (NO2)	8	20
Carbon Monoxide (CO)	11	26
Sulfur Dioxide (SO2)	2	17
Total Suspended		
Particulate Matter (TSP)	0	27
Sulfate (SO4)	0	7

^{*}For valid stations, see Attachments C and D.

The criteria used to select a <u>valid</u> station for gaseous pollutants are as follows:

- 1. The station must report air quality data for at least 50% of the total number of hours per month to be a <u>valid</u> station for that month.
- 2. The station must have at least 7 <u>valid</u> months per year to be a <u>valid</u> station for that year.
- 3. The station must be <u>valid</u> in 5 out of 7 years during the 1st period and in 3 out of 4 years during the 2nd period to be a <u>valid</u> station.

The criteria used to select a <u>valid</u> station for a particulate pollutant are similar to those for a gaseous pollutant. However, the first criterion is changed to the following:

1. A station must report air quality data for at least two days per month to be $\underline{\text{valid}}$ for that month.

<u>Valid</u> stations selected according to the criteria described above are listed in Attachment C for the first period, 1966-1972, and in Attachment D for the second period, 1973-1976.

2.2 Development of Air Quality Data Base

When a human exposure data base is to be developed from ambient air quality monitoring data, a number of problems arise: First, the number of monitoring stations that report an adequate number of hourly or daily air quality values varies from year to year. If a changing set of monitoring stations is used to estimate human exposure levels at a non-station location, the reliability and accuracy of the estimated exposure values may vary from year to year. Therefore, the same set of valid monitoring stations was used to compute human exposure levels at

each postal zip-code area for each pollutant and for each of the two study periods, 1966-1972 and 1973-1976.

Second, air quality data for a gaseous pollutant are too voluminous for simple computation of monthly exposure values. Therefore, the gaseous pollutant data were compressed by computing a cumulative distribution of hourly concentrations for each month at each station. The cumulative distribution was approximated by computing concentration values at selected percentiles: every .25% above 99% (four values), every 1% above 90% (nine), every 2.5% above 70% (eight), every 5% above 50% (four), every 10% equal to or above 0% (six). A concentration value at the 100th percentile was set equal to the maximum measured concentration, while that at the 0th percentile was set equal to the minimum measured concentration. In addition to the above percentile values, an arithmetic mean concentration was also computed each month for each station. By converting hourly concentration data into the percentile form described above, we compressed the air monitoring data by about 22 to 1 (see Attachment G).

To compute the monthly percentile statistics, the original hourly concentration readings were sorted in a descending order. The highest concentration was ranked as r=1, the second highest as r=2, and so forth. For gaseous pollutant data, the percentile, P, of each concentration value was computed by the following approximate formula:

$$P = 1 - \frac{r}{N+1}$$
 (1)

where ${\bf r}$ is the rank of that concentration value and N is the sample size of the gaseous pollutant concentration data for a given month and station.

Third, air quality data for a particulate pollutant are too few to construct a believable cumulative distribution. While the cumulative distribution is to be specified by concentration values at the $31 \ (4+9+8+4+6)$ selected percentiles, the total number of particulate concentrations reported each month by each monitor seldom

exceeds 6 (30 days/every 5th day). Therefore, the particulate concentration data are listed as they are, only by removing the dates on which the samples were taken (see Attachments H and I). In addition to the raw concentration data, the number of samples per month, the maximum and minimum values, and the geometric mean and standard deviation were computed for each month and station.

2.3 Exposure Data Base at Monitoring Sites

From the air quality data compiled each month for each station, monthly statistics of various exposure parameters were computed. The exposure parameters used in the current data base are as follows:

Symbol .	Exposure Parameter	Unit
MEAN	Mean Exposure Level (a mean pollutant concentration over	
	a month)	[Conc]
DOSE	Total Exposre (a cumulative	
	exposure to air pollution	
	over a month)	[Conc-Hr]
F	Exceedance Frequency (the	
	number of hours exposed to	
	concentrations above a given	
	threshold over a month)	[Hr]
E	Excess Exposure (a cumulative	
	exposure to concentrations	
	above a given threshold over	
	a month)	[Conc-Hr]

The mean exposure level (MEAN) at a monitoring site is approximated by a mean concentration level measured at that site. To make it consistent with the air quality standards, the mean exposure level is given by an arithmetic mean of hourly concentrations for the gaseous pollutants (OX, NO2, SO2 and CO) and by a geometric mean of 24-hour concentrations for the particulate pollutants (TSP and SO4).

The total exposure (DOSE) is given by a time integral of concentrations over a month. The total exposure can also be computed by multiplying the mean exposure level by the number of hours in a given month, i.e.,

where T is the number of hours in that month.

A mean exposure level and a total exposure have been used for various epidemiological studies [Shy et al. 1970; Lave and Seskin 1977] to indicate a different level of human exposure for each sub-group of the study population. The two exposure parameters, DOSE and MEAN, may be a good exposure indicator if adverse air pollution health effects occur progressively from very low exposures to high exposures. However, an air quality standard has been set based on the premise that there exists a threshold concentration below which the public may never experience any adverse health effect from exposures to air pollution.

To facilitate future investigations of the above premise, the current human exposure data base includes two additional exposure parameters: Exceedance Frequency, F, and Excess Exposure, E. These parameter values are computed for each of the four threshold levels, which are listed in Attachment B for each of the six pollutants under study. As illustrated in Figure 3, the exceedance frequency and the excess exposure are given, respectively, by

$$F = T \quad x \quad P \tag{3}$$

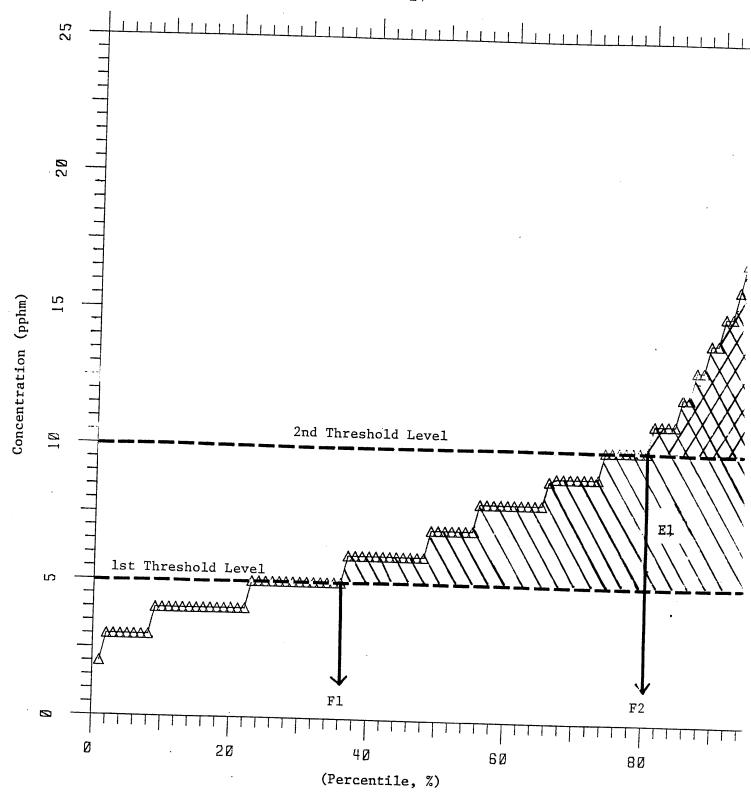


Figure 3. Cumulative Distribution of Hourly NO₂ Concentrations at Downtown Los Angeles for January 1978. (Excess exposures and exceedance frequencies for the 1st and 2nd threshold levels are given, respectively, by El=hatched area x 744 hr, E2=double hatched area x 744 hr, F1=.36 x 744 hr, and F2=.80 x 744 hr.)

$$E = T \times \int_{P_{O}}^{100} [C(P) - C_{S}] dP$$
 (4)

where P is the percentile at which a percentile concentration, C(P) becomes equal to a given threshold level, C_S .

Threshold Levels

For the last two exposure parameters, E and F, the parameter values are computed for each of four selected threshold levels as listed in Attachment B. For each pollutant, the threshold levels were chosen by considering actually occurring ambient levels and established concentration levels such as those used in the air quality standard and the health alert level. For example, an ambient level of daily maximum 1-hour oxidant concentrations varies from about 5 pphm to 50 pphm in the South Coast Air Basin (SCAB). The established oxidant concentration levels are 8 pphm (old national standard), 10 pphm (California standard), 12 pphm (new national standard), 20 pphm (lst stage alert level), and 35 pphm (2nd stage alert level). Therefore, we select the following threshold levels: 10 pphm for the lst level; 15 pphm for the 2nd level; 20 pphm for the 3rd level; and 25 pphm for the 4th level.

Similarly, the thresholds for NO2 were selected as 5 pphm (national primary annual standard), 15 pphm, 20 pphm, and 25 pphm (California 1-hr standard). These threshold levels are well within the normal range of daily maximum 1-hr NO2 concentrations occurring in the SCAB region. The thresholds chosen for SO2 are 2, 4 (California 24-hr standard), 8, and 14 pphm (national primary 24-hr standard); those for CO are 5, 10 (California 12-hr standard), 15, and 20 pphm.

For TSP, there are California and national standards for both 24-hr concentrations and annual geometric mean concentrations. Therefore, the thresholds for TSP were mostly selected from the levels designated by those standards: 60 (California annual standard), 100 (California 24-hr standard), 150 (national secondary annual standard) and $200\mu g/m^3$. On the other hand, there is only one air quality standard for SO4. This California 24-hr standard (25 $\mu g/m^3$) is violated rather

infrequently in the SCAB region. Therefore, the threshold levels were chosen as 10, 15, 20, and 25 $\mu g/m^3$

Filling in Missing Data

Although rather rigid criteria were employed to select <u>valid</u> stations for this data base development, some <u>valid</u> stations had several <u>invalid</u> months, i.e., months with less-than-complete data during each of the two study periods: January 1966 through December 1972, and January 1973 through December 1976. In estimating values of each exposure parameter for those <u>invalid</u> months, we tried three different methods of filling in the missing values.

Determine a typical seasonal pattern from J years
 of good data and then estimate a missing monthly
 value, say, the ith monthly average in the jth year,
 X
 i,j

$$X_{i,j} = M_{i,j} Y_{j}$$
(5)

where M $_{\rm i}$ is the adjustment factor for the ith month and Y $_{\rm j}$ annual average of the jth year.

ii) Fill in the missing values by averaging the corresponding monthly values in two adjacent years, i.e., take

$$X_{i,j} = (X_{i,j-1} + X_{i,j+1}) / 2$$
 (6)

iii) Fill in the missing value by substituting with the corresponding monthly value in one adjacent year, i.e., take

or
$$X_{i,j} = X_{i,j-1}$$

$$X_{i,j} = X_{i,j+1}$$
(7)

Initially we thought that the first method would yield a better estimate for a missing monthly value than the second and third methods. However, our trial results of each estimation method have indicated that the second and third methods often yield a better estimate for the missing monthly values than does the first method. In this exercise, we treated some valid months as missing months and compared the estimated values with the actual values to see the performances of the three estimation methods.

The above finding led to a more rigorous comparison of the three estimation methods. As described in detail in Attachment Q, the theoretical comparison also showed that the second and third methods were superior to the first method when a year had 5 or more missing months.

The air quality data base used for this study exhibited the following characteristics: If a station has a bad year, the station has many missing months in that year. If a station has a good (i.e., valid) year, the station seldom has a missing month in that year.

Because of the empirical and theoretical results and the missing data characteristics, we decided to use a combination of the second and third methods, i.e., the second method for an intermediate year, and the third method for either the beginning year or the ending year of each of the two periods, 1966-1972, and 1973-1976.

2.4 Exposure Data Base at Postal Zip-Code Areas

Human exposure to air pollution at each postal zip-code area is estimated by applying a spatial interpolation formula to the exposure values at air monitoring sites. After several trials of applying various spatial interpolation schemes to air quality monitoring data, we decided to use the spatial interpolation program developed by Guardado and Sommers [1977] at the Pacific Southwest Forest and Range Experimental Station, Berkeley, California. The program includes the following:

- i) An unevenly spaced orthogonal grid is developed such that each row and column passes through one of the air quality monitoring sites. This grid is called an initial "guess field" grid.
- ii) An exposure parameter value at an initial "guess field" grid point is estimated by an inverse square-of-distance weighting interpolation formula applied to the known exposure values at the monitoring sites, i.e.,

$$X = \frac{\sum_{i=1}^{n} w_i X_i}{\sum_{i=1}^{n} w_i}$$

and

$$w_{i} = \frac{1}{d_{i}^{2}}$$

where X is the interpolated value at the initial "guess field" grid point, X_{i} are the known exposure values at the monitoring sites, and d_{i} is the distance between the interpolation point and the ith monitoring site (i = 1, 2, ..., n).

- iii) The initial guess field rows and columns are then adjusted by using the known exposure values and a parabolic leapfrogging technique, which is discussed in Attachment P.
 - iv) Finally, an exposure value at each postal zip-code area is estimated by a cubic spline interpolation method applied to the corrected field rows and columns.

In actual interpolation, we assigned the smallest measured station value to the extreme northwest, northeast, and southeast corners of the study area. As Figure 1 shows, the study area is bounded by the coastal line in the southeast direction. Therefore, we did not assign any boundary value to the extreme southwest corner. These boundary values were used to force all the interpolated values to be within the range of the measured station values. In addition, the interpolated values at each postal zip-code area are indicated by a confidence level, which was determined by comparing a typical representative distance around a monitor with a distance between the centroid of that zip-code area and the nearest valid monitoring station.

Table 2 presents the typical representative distance around a monitor for each of the six study pollutants. If the distance of the nearest monitoring station is less than or equal to the typical representative distance, the exposure data at that postal zip-code area are designated as "A", the highest confidence level. If the interdistance is in between the typical representative distance and twice that distance, the exposure data are designated as "B", the moderate confidence level. If the inter-distance exceeds twice the typical representative distance, the data are designated as "C", the lowest confidence level.

Table 2. Typical "Representative Area" of a Monitoring Station Estimated for Each Major Pollutant from the Expected Pollution Gradients of Long-Term Concentration Statistics [After U.S. EPA 1977]

Pollutant	Represent Distance fr	ative Area om a Station
OX	10 miles	16 km
NO2	. 5	8
SO2	3	4.8
СО	1	1.6
TSP	3	4.8
S04	20	32

The distributions of postal zip-code areas over the study area are plotted in Attachments E and F by using the three confidence levels, A, B, and C. As seen from Attachments El and Fl for photochemical oxidants, OX, the great majority of the zip-code areas are classed as A or B, indicating that the exposure values at those zip-code areas are at least believable. On the other hand, Attachments E3 and F3 for carbon monoxide (CO) are filled with the letter "C", indicating that the exposure values are not more than a mere guess. The confidence in the exposure data for other pollutants falls in between those of the above two pollutants (see Attachments E2, E4, F2, F4, F5 and F6).

3. RECOMMENDED DATA BASE IMPROVEMENTS

The human exposure data base developed under the present study will provide researchers of air pollution health effects with a convenient and reliable data source for estimating exposures of individual members of a study population to air pollution from various pollutants. The data base includes monthly exposure data of several parameters for each of the four gaseous (OX, NO2, SO2 and CO) and two particulate (TSP and SO4) pollutants over the entire South Coast Air Basin during the eleven-year period, January 1966 through December 1976.

The exposure parameters include the mean exposure level, the total exposure, and the exceedance frequency and excess exposure whose parameter values are computed for each of four different threshold levels selected for each pollutant. Values of these exposure parameters are given for each of some 400 postal zip-code areas scattered over the study region as well as at each of the air monitoring stations whose air quality data were used to develop data base.

While the present data base is more complete than and superior to those developed by earlier epidemiological studies, the data base needs to be updated periodically and further refined in some exposure estimates:

- By now, the 1977 and 1978 air quality data have become available and the 1979 air quality data will soon become available. These recent air quality data can be used to update the human exposure data base.
- In the development of the present data base, a purely
 mathematical interpolation scheme was used to estimate
 exposure values at each postal zip-code area from those
 computed at air monitoring sites. However, if emissions and
 meteorological data are also employed in such an estimation

process, one may be able to improve exposure estimates, particularly at those postal zip-code areas distant from any air monitoring station.

• A person's work place as well as his residence location can be found through simple survey questionnaires. Then the person's exposure will be better estimated from the sum of the exposure value at his work place and that at his residence than from the latter alone. To facilitate such refinement of exposure estimates, exposure during the typical work hours (say, weekdays 7 A.M. to 5 P.M.) and that during the non-work hours may need to be computed in addition to the present "all day exposure" estimates.

4. REFERENCES

- Guardado, J.L. and W.T. Sommers, "Interpolation of Unevenly Spaced Data Using a Parabolic Leapfrog Correction Method and Cubic Splines," USDA Forest Service Research Note PSW-324, Pacific Southwest Forest and Range Experiment Station, Berkeley, California, 1977.
- 2. Horie, Y., J. Cassmassi, L. Lai and L. Gurtowski, "Weekend/Weekday Differences in Oxidants and Their Precursors," EPA-450/4-79-013, U.S. Environmental Protection Agency/OAQPS, Research Triangle Park, North Carolina, March 1979.
- 3. Lave, L.B. and E.P. Seskin, <u>Air Pollution and Human Health</u>, Johns Hopkins University Press, 1977.
- 4. Shy, C.M., et al., "The Chattanooga School Children Study: Effects of Community Exposure to Nitrogen Dioxide," <u>J. APCA</u>, Vol. 20, No. 8, August 1970, pp. 539-545.
- 5. U.S. EPA, "Guideline on Procedures for Constructing Air Pollution Isopleth Profiles and Population Exposure Analysis," EPA-450/2-77-024a (OAQPS No. 1.2-083), U.S. Environmental Protection Agency/OAQPS, Research Triangle Park, North Carolina, October 1977.

Attachments

DESCRIPTION AND SAMPLES OF HUMAN EXPOSURE DATA BASE

Description and Samples of Human Exposure Data Base

Attachment	Data Base Description
Α	Data base documentation
B .	Concentration threshold levels used for the human exposure data base
C	Monitoring stations and data availability during 1966-1972
D	Monitoring stations and data availability during 1973-1976
E	Postal zip code area classified according to the distance from the nearest monitoring station during 1966-1972
F	Postal zip code area classified according to the distance from the nearest monitoring station during 1973-1976
	Data Base Samples for Downtown L.A. Site
G	Gaseous Pollutant Percentile Data
Н	Particulate Air Quality Data Base
I	Sulfate Air Quality Data
J	Oxidant Exposure Parameter Data Base
K	NO ₂ Exposure Parameter Data Base
L	CO Exposure Parameter Data Base
М	SO ₂ Exposure Parameter Data Base
N	TSP Exposure Parameter Data Base
0	Sulfate Exposure Parameter Data Base
Р	Cubic Spline Interpolation Scheme With A Parabolic Leapfrog Correction
Q	Comparing Methods of Filling In Missing Data

FIELD

HUMAN EXPOSURE DATA BASE

ALL OF THE HUMAN EXPOSURE DATA BASES HAVE THE SAME STYLE HEADER IN THE FORMAT BELOW:

COLUMNS COMMENTS

```
1-3 CO1,NO2,OXI,SO2,TSP,SO4
     VARIABLE
                            4 A,B,C (SEE METHOD CODE TABLE)
5-9 (SEE STATION NUMBER AND ZIP TABLES)
    METHOD CODE
2.
     STATION NUMBER OR ZIP
                              10-11 66-76
 4.
     MONTH
                              12-13 1-12
      THE EXPOSURE PARAMETER DATA BASE HAS THE FORMAT: (See Attachments J thru M)
 1-5. STANDARD HEADER
                                1-13
                               14-15 1-6 (1=HIGHEST CONFIDENCE)
 6. CONFIDENCE INDICATOR
 7. MEAN
                               16-20 (F5.1)
 8. DOSAGE (CONC.HRS)
                               21-26 (16)
                               27-32 (16)
9. EXCESS FREQUENCY (HRS)
               > LEVEL 1
                                       (SEE EXPOSURE LEVEL TABLE)
 10. EXCESS DOSAGE(CONC.HRS) 33-38 (16)
               > LEVEL 1
 11. EXCESS FREQUENCY (HRS)
                               39-44 (16)
                                       (SEE EXPOSURE LEVEL TABLE)
               > LEVEL 2
 12. EXCESS DOSAGE(CONC.HRS) 45-50 (16)
               > LEVEL 2
 13. EXCESS FREQUENCY (HRS)
                               51-56 (16)
                                       (SEE EXPOSURE LEVEL TABLE)
               > LEVEL 3
 14. EXCESS DOSAGE(CONC.HRS) 57-62 (16)
               > LEVEL 3
 15. EXCESS FREQUENCY (HRS)
                               63-68 (16)
                                       (SEE EXPOSURE LEVEL TABLE)
               > LEVEL 4
 16. EXCESS DOSAGE(CONC.HRS) 69-74 (I6)
               > LEVEL 4
1
     THE PERCENTILE DATA BASE IS IN THE FORMAT: (See Attachment H)
 1-5. STANDARD HEADER
                             1-13
      CONFIDENCE INDICATOR 14-15 1-6 (1=HIGHEST CONFIDENCE)
 6.
 7.
      MEAN
                           16-20 (F5.1)
 08. PERCENTILE 1.000
                             021-023
09. PERCENTILE .9975
10. PERCENTILE .9950
11. PERCENTILE .9925
                             024-026
                             027-029
                             030-032
 12.
     PERCENTILE .9900
                             033-035
 13. PERCENTILE .9800
                             036-038
                             039-041
 14. PERCENTILE .9700
 15. PERCENTILE .9600
                             042-044
 16. PERCENTILE .9500
                             045-047
     PERCENTILE .9400
 17.
                             048-050
     PERCENTILE .9300
 18.
                             051-053
 19. PERCENTILE .9200
                             054-056
 20. PERCENTILE .9100
                             057-059
21. PERCENTILE .9000
22. PERCENTILE .8750
23. PERCENTILE .8500
                             060-062
                             063-065
                             830-330
 24. PERCENTILE .8250
                             069-071
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Attachment A

Data Base Documentation

```
PERCENTILE .8000
PERCENTILE .7750
PERCENTILE .7500
25.
                               072-074
26.
                               075-077
27.
                               078-080
      PERCENTILE .7250
28.
                               081-083
      PERCENTILE .7000
29.
                               084-086
      PERCENTILE .6500
30.
                               087-089
31.
      PERCENTILE .6000
                               090-092
      PERCENTILE .5500
32.
                               093-095
33.
     PERCENTILE .5000
                               096-098
34.
     PERCENTILE .4000
                               099-101
35.
     PERCENTILE .3000
                               102-104
     PERCENTILE .2000
36.
                               105-107
     PERCENTILE .1000
PERCENTILE .0000
37.
                               108-110
38.
                             111-113
```

THE SO4 AND TSP DATA BASES DO NOT FOLLOW THE PERCENTILE DATA BASE FORMAT BECUASE OF THE LIMITED NUMBER OF MEASUREMENTS (5-6 PER MONTH AVERAGE VERSUS 720 PER MONTH AVERAGE). THE SO4 AND TSP DATA BASES HAVE THE FORMAT:

```
(See Attachments H and I)
1-5. STANDARD HEADER
                             1-13
6. N - NUMBER OF CASES
                            14-15
                                   (12) ( <= 15 )
    MAX - MAXIMUM VALUE
                            16-21
                                   (F6.1)
    MIN - MINIMUM VALUE
                            22-27
                                   (F6.1)
    GMEAN-GEOMETRIC MEAN
                            28-33
                                   (F6.1)
10. G.SD-GEOMETRIC
                            34-39
                                   (F6.1)
    STANDARD DEVIATION
11. VALUES-
                            40-130
      ACTUAL MEASUREMENTS
                                  (15F6.1)
```

Attachment A (Continued)

Data Base Documentation

2		Co	ncentration Thr	reshold	
	Pollutant	1st Level	2nd Level	3rd Level	4th Level
	ox	10 pphm (Calif.std.)	15 pphm	20 pphm (1st stage alert)	25 pphm
l hr Concentration —	NO ₂	5 pphm (primary 1 yr. std.)	15 pphm	20 pphm	25 pphm (Calif. 1-Hr. std.)
Concer	S0 ₂	2 pphm	4 pphm (Calif. 24-Hr. std.)	8 pphm	14 pphm (primary 24-Hr. std.)
<u> </u>	CO	5 ppm	10 ppm (Calif. 12-Hr. std.)	15 ppm	20 ppm
24 hr entration —	TSP ,	60 ug/m ³ (Calif. l yr. std.)	100 ug/m ³ (Calif. 24-Hr. std.)	150 ug/m ³ (secondary 1 yr. std.)	200 ug/m ³
24 hr Concentration	S0 ₄	10 ug/m ³	15 ug/m ³	20 ug/m ³	25 ug/m ³ (Calif. 24-Hr. std.)

Attachment B

Concentration Threshold Levels Used For

Human Exposure Data Base

N AAAAAAAAAA
_
0 4 4 4
0 A A A A A A A A A A A A A A A A A A A
0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0 0
0 V AA AAAAAAA V AA
PAPAPAP PA 6
ANAHEIM SAN. BERNARDINO REDLANDS LOS. ANGELES. DOWNTOWN AZUSA BURBANK WEST. LOS. ANGELES LONG. BEACH RESEDA POMONA LENNOX
30176 36151 36165 70001 70069 70071 70072 70074 70075
OXIDANT OXIDANT OXIDANT OXIDANT OXIDANT OXIDANT OXIDANT OXIDANT

MEASUREMENT METHOD

OXIDANT A=12. =Total Oxidant/Colorimetric B=36. =0zone/UV Photometric C=36.01 =0zone/Chemiluminescent

Attachment C1. Monitoring Stations Reporting Adequate Amount of Valid Oxidant Data During 1966-1972.

i.i.
0
Z AAAAAA
02
6044444444
84444444
2
9 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4
LOS. ANGELES. DOWNTOWN AZUSA BURBANK WEST. LOS. ANGELES LONG. BEACH RESEDA POMONA LENNOX
70001 1 70069 6 70073 6 70072 L 70074 F 70076 F
ROGEN DIOXIDE 7

NITROGEN DIOXIDE A=21. =Colorimetric (Saltzman)
B=21.01 =Chemiluminescent

Attachment C2. Monitoring Stations Reporting Adequate Amount of Valid NO $_{
m 2}$ Data During 1966-1972.

1
7 4444444444
_
70 A A A
0 444444444 0
9 AAAAAAAA 8
0 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4
9 44 44444 44 00
ANAHEIM SAN. BERNARDIND REDLANDS LOS. ANGELES. DOWNTOWN AZUSA BURBANK WEST. LOS. ANGELES LONG. BEACH RESEDA POMONA LENNOX
30176 36151 36165 70001 70069 70071 70072 70072 70074
N MONOXIDE

CARBON L MEASUREMENT METHOD

Carbon Monoxide A=15. =Long Path NDIR B=15.0T =FID

69 A ANAHEIM LOS ANGELES DOWNTOWN 30176 70001 SULFUR DIOXIDE SULFUR DIOXIDE

68 A 67 A 66 A A

MEASUREMENT METHOD

SULFUR DIOXIDE

A=18. =COULOMETRIC B=18.01=FLAME PHOTOMETRIC

Attachment C4. Monitoring Stations Reporting Adequate Amount of Valid SO2 Data During 1966-1972,

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Σ.	3RA	COSTA.MESA.HARBOR		LOS.ALAMITOS.ORANGEMOON	INDIO.OASIS.ST	HEMET.STATE.ST	RIVERSIDE, RUBIDOUX	RIVERSIDE, MAGNOLIA		SAN. BERNARDIND	SQ	CHINO, RIVERSIDE, AVE	CIVIC CENTER	ABB	FONTANA, FOOTHILL		PAULA	ES, DOWNTOWN			. ANGELES							MALNUT		
6 ANAHEIM		_											_		•				⋖	ш	WEST.LOS			<u></u>		WHITTIER	NEMHALI	PASADENA.	LYNWOOL	
30176	3017	30185	3018	3019	3313	3314	3314	33146	33149	36151	36165	36173	3617	36175	36176	56402	56404	70001	70060	70069	70071	70072	70074	70075	70076	70080	70081	70083	/0084	
OXIDANT	· OXIDANT	OXIDANT	UXIDANI	UXIDANT	UXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	DXIDANT	DXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	OXIDANT	DXIDANT	UXIDANT	OXIDANT	OXIDANT	UXIDANT	OXIDANT	OXIDANT	OXIDANT	UXIDANI	OVIDANI	

A=12.__=TOTAL OXIDANT/COLORIMETRIC B=36.__=OZONE/UV PHOTOMETRIC C=36.01=OZONE/CHEMILUMINESCENT

OXIDANT

Attachment D1. Monitoring Stations Reporting Adequate Amount of Valid Oxidant Data During 1973-1976.

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73 A	∢ ∢	∢ ∢	< <	∢ ⊲	< <	<	₹	. ⊲	< <	< <	€ <	ζ.	∢ •	∢.	⋖	
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30176 30177 30185	33139	33146	36151 36165	70001	70060	70063	/00/1	70072	70074	70075	2002	70080	70081	70083	70084	1000
NI TROGEN, DI OXI DE MI TROGEN, DI OXI DE TROGEN, DI OXI DE	NITROGEN. DIOXIDE	NI TROCEN. DIOXIDE	NITROGEN.DIOXIDE	NI TROGEN. DIOXIDE	NI TROGEN, DIOXIDE			NI IRUGEN. DI UXI DE		IOI	DIOXI	DIOXI	NI TROGEN, DI OXI DE	NI TROGEN, DIOXIDE	ROGEN, DIOXI	

NITROGEN DIOXIDE A=21.__=COLORIMETRIC(SALIZMAN)
B=21.01=CHEMILUMINESCENT

Attachment D2. Monitoring Stations Reporting Adequate Amount of Valid NO₂ Data During 1973-1976.

74 75 76	A A A				V	⋖	. Φ		: ∢	: ∢	< ⊲	< ⊲		τ <	τ <	₹•	V •	∢ <	₹ •	€ •	∢ •	¥	¥	۷	⋖	⋖	¥
73	V	V	V		∢		⋖		∢	¥	< ∢	. ⊲	. ⊲	ζ		ζ <	∢ <	€ <	Σ <	τ <	∢•	∢.	⋖	∢	⋖	∢	
	ANAHEIM	LA.HAB			INDIO. DASI	NORCO. PRADO. PARK	RIVERSIDE, RUBIDOUX	RIVERSIDE.MAGNOLIA	SAN. BERNARDINO		CHINO.RIVERSIDE.AVE	UPLAND.CIVIC.CENTER	UPLAND, ARB	F		A711SA	BURBANK	WEST, LOS, ANGEL ES	L ONG. BEACH	RESERIA	DOMONA					PASADENA, WALNU!	LYNWOUD
	30176	301/	30185	30189	33138	٠,٠,٠	(,)	י ניין	(*)			36174								1	. 1	- 1	- 1	70007	10007	70007	/0084
	CARBON. MONOXIDE	CARBOIN, FIGNOX I DE	CARBON, FIONOX DE	CARBON FIONOXIDE		CARBON, FIUNUX I DE	CAHBON. MONOXIDE	CARBON, MONOXIDE	CARBON. MONOXIDE	CARBON. MONOXIDE	CARBON.MONOXIDE	CARBON. MONOXIDE	CARBON. MONOXIDE	CARBON. MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CARBON, MONOXIDE	CABBON MONOXIDE	CARRON MONIOVIDE	CADBON MONOVIDE	TATACINON MODIA	CALBON HONOXIDE	CARBON, FIGNOX I DE

CARBON MONOXIDE A=15. _=LONG PATH NDIR B=15.01=FID Attachment D3. Monitoring Stations Reporting Adequate Amount of Valid CO Data During 1973-1976.

במינאסנם מוים וויס			73	74	75	76
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SULFUR. DIOXIDE		COSTA, MESA, HARBOR	: <	< <	c	۲ <
SULFUR, DIOXIDE		I DS. ALAMITOS DRANGELIODO	c	τ.		∢ ·
SIII FIIR, DIOXIDE		CAN DEDNADOTAD	•	Œ.	⋖	∢
SIL FIR DIXIDE		CANOTINATION OF THE PROPERTY O	∢ •	⋖	⋖	¥
	-	LOG. AIVORLED. DOMINI UMIN	∢	ď	⋖	¥
	-	AZUSA	⋖	V	¥	۷
SULFUR. DIUXIUE		BURBANK	۵	٧	: <	: <
SULFUR.DIOXIDE	•	WEST, LOS, ANGFI FS	<	< <	τ <	۲ -
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	•		⋖	ď	⋖	⋖
SULFUR. DIUXIUE	-	KESEDA	4	٧	<	: <
SULFUR.DIOXIDE	•	POMONA	< <	< <	٠ -	۲.
SULFUR, DIOXIDE	•	FINDS	₹ •	₹ .	₹ -	∢
	•		¥	∢	⋖	∢
SULT UN. DI UAI DE	-	MILLIER	A	۵	۷	<
SULFUR DIOXIDE	•	NFIZHAL I	<	: •	٠ ،	ς .
SUI FUR DIOXIDE		DACADENA 1(A) NI F	∢ •	₹ '	⋖	¥
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						:

MEASUREMENT METHOD

SULFUR DIOXIDE A=18. =COULOMETRIC B=18.01=FLAME PHOTOMETRIC

Monitoring Stations Reporting Adequate Amount of Valid ${
m SO}_2$ Data During 1973-1976. Attachment D4.

TSP 56412 PORT HUENEME TSP 56413 SIMI VALLEY TSP 56414 VENTURA TELEGRAPH RD TSP 56415 THOUSAND OAKS WINDSOR TSP 70001 LOS ANGELES DOWNTOWN TSP 70060 AZUSA TSP 70076 LENNOX	TSP TSP TSP TSP TSP TSP	SP 30177 SP 30184 SP 30185 SP 30186 SP 30189 SP 30190 SP 33126 SP 33144 SP 33144 SP 36151 P 56401 P 56402 P 56408 P 56409 P 56410 P 56412 P 56413 P 56415 P 56415 P 56416	SIMI VALLEY VENTURA TELEGRAPH RD THOUSAND OAKS WINDSOR LOS ANGELES DOWNTOWN AZUSA	A A A A A	A A A A A	A A A A	A A A A
		70076	LENNOX	Α	Α	Α	A A

MEASUREMENT METHOD

TSP A = HIGH VOLUME SAMPLING

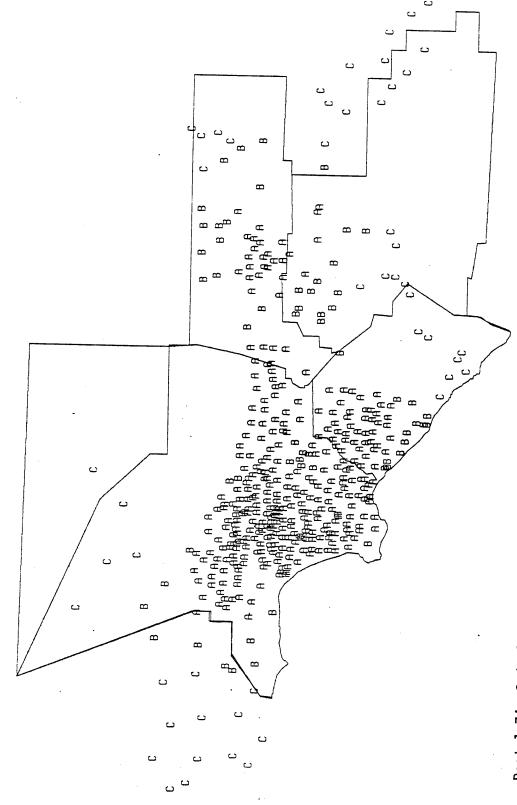
Attachment D5. Monitoring Stations Reporting Adequate Amount of Valid TSP Data During 1973-1976.

			<u>73</u>	<u>74</u>	<u>75</u>	<u>76</u>
S04	70001	LOS ANGELES DOWNTOW	У	Α	Α	۸
S04	70060	AZUSA	Ä	Λ	Α	A
S04	70071	WEST LOS ANGELES	· · · · · · · · · · · · · · · · · · ·	A .	A	A
S04	70074	RESEDA	A	A	A	, А
S04	70074	LENNOX	A	Α	А	Α
S04			А	Α	Α	Α
	70083	PASADENA WALNUT	А	Α	Α	Α
S04	70084	LYNWOOD	А	Α	Α	A

MEASUREMENT METHOD

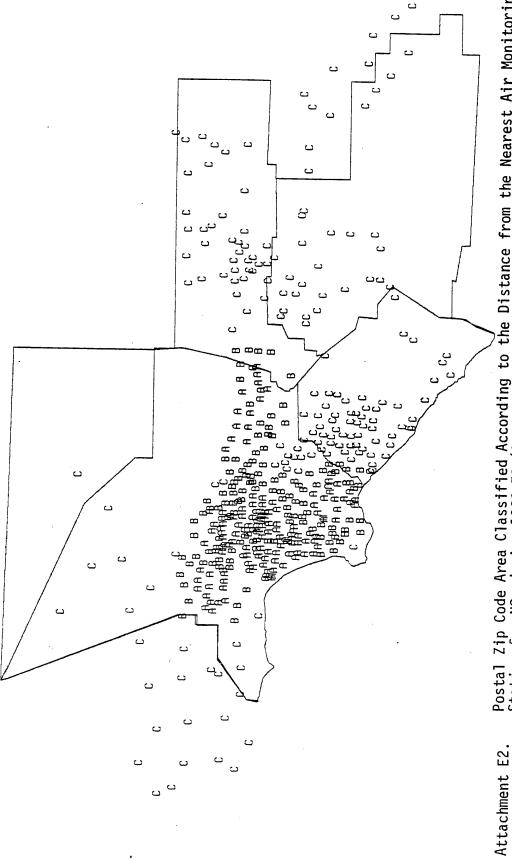
SULFATE A = HIGH VOLUME SAMPLING

Attachment D6. Monitoring Stations Reporting Adequate Amount of Valid Sulfate Data During 1973-1976.

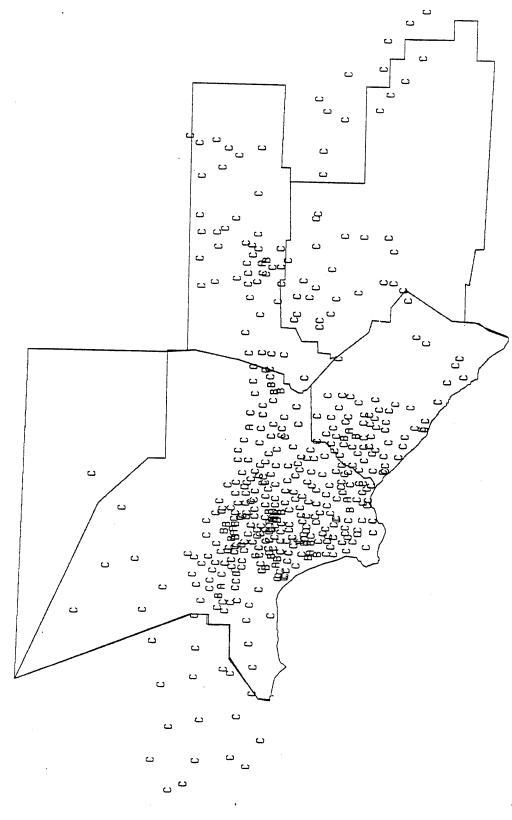


Postal Zip Code Area Classified According to the Distance From the Nearest Air Monitoring Station for Oxidant during 1966-72 (A \leq 16 Km₂16 Km < B \leq 32 Km₂C > 32 Km).

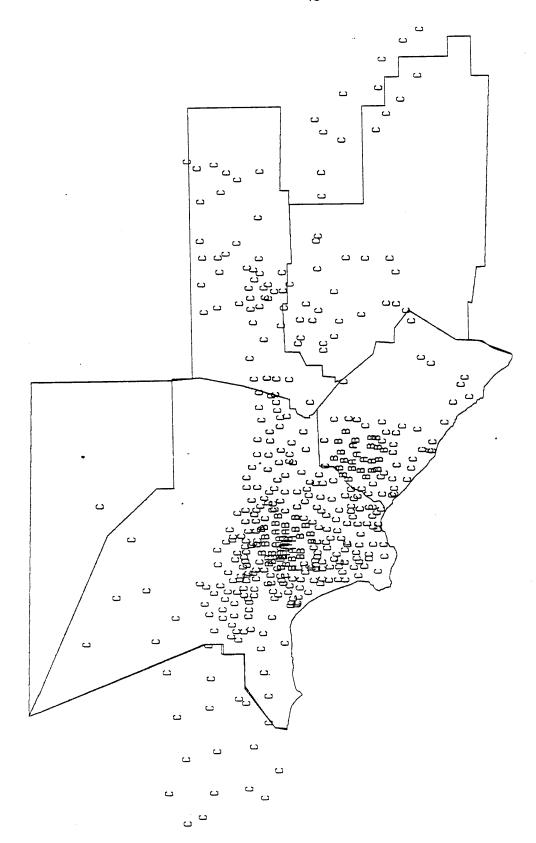
Attachment El



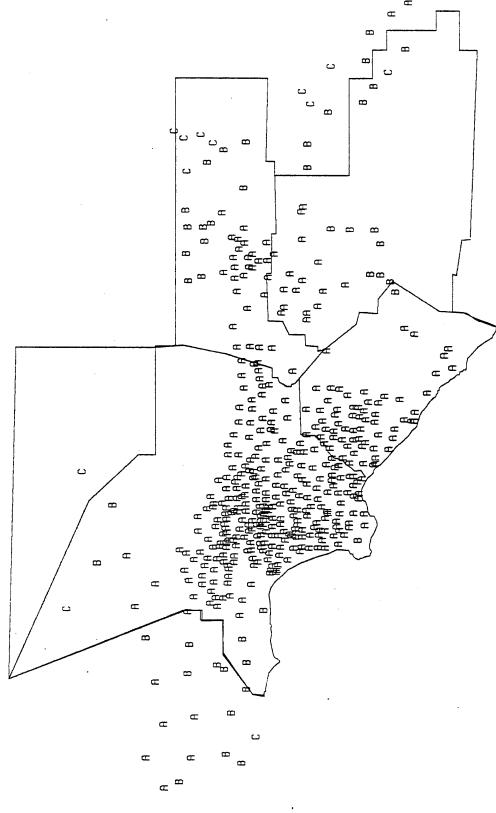
Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for N0₂ during 1966-72 (A \leq 8 Km,8 Km < B \leq 16 Km, C > 16 Km).



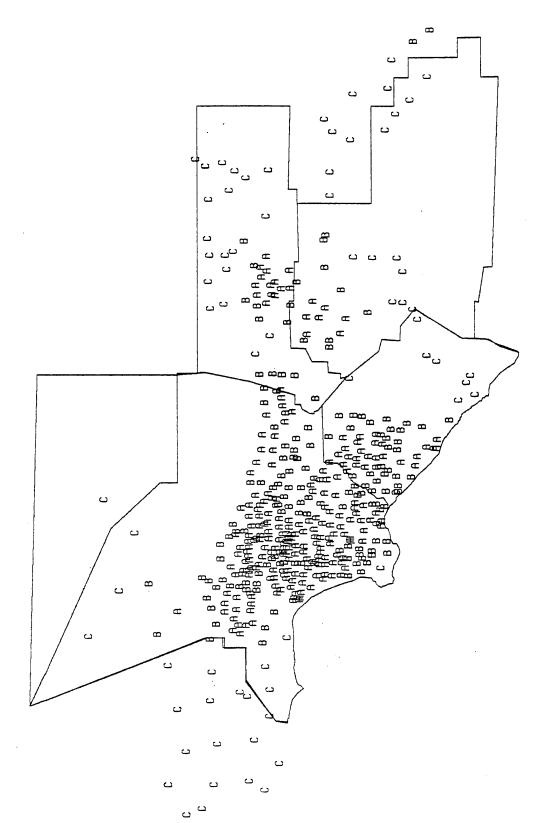
Postal Zip Code Area Classified According to the Distance From the Nearest Air Monitoring Station for CO during 1966-72 (A \leq 1.6 Km,1.6 Km < B \leq 3.2 Km,C > 3.2 Km). Attachment E3



Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for $\rm SO_2$ during 1966-72 (A < 4.8 Km,4.8 Km < B < 9.6 Km,C > 9.6 Km). Attachment E4

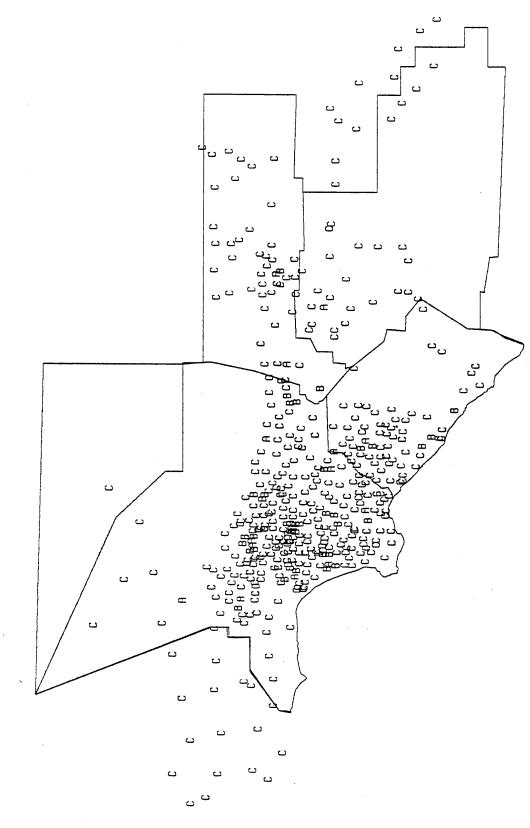


Attachment F1 Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for Oxidant during 1973-76 (A \leq 16 Km,32 Km < B \leq 32 Km,C > 32 Km).

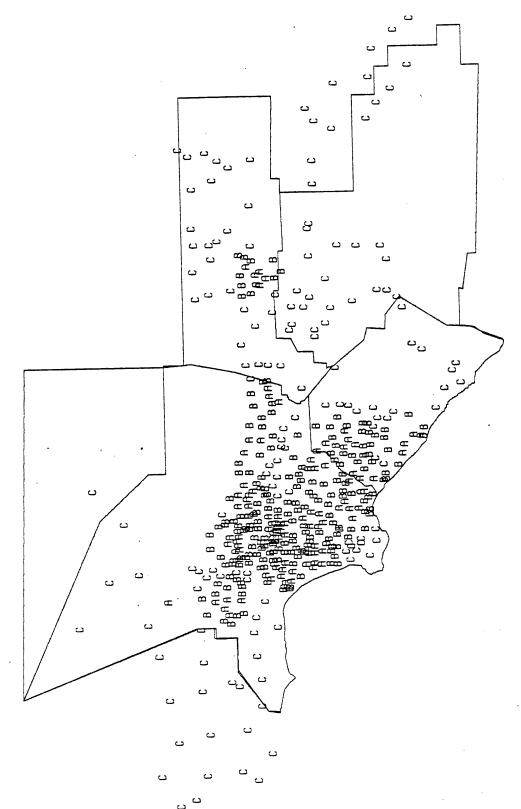


Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for NO₂ during 1973-76 (A \leq 8 Km, 8 Km < B \leq 16 Km, C > 16 Km).

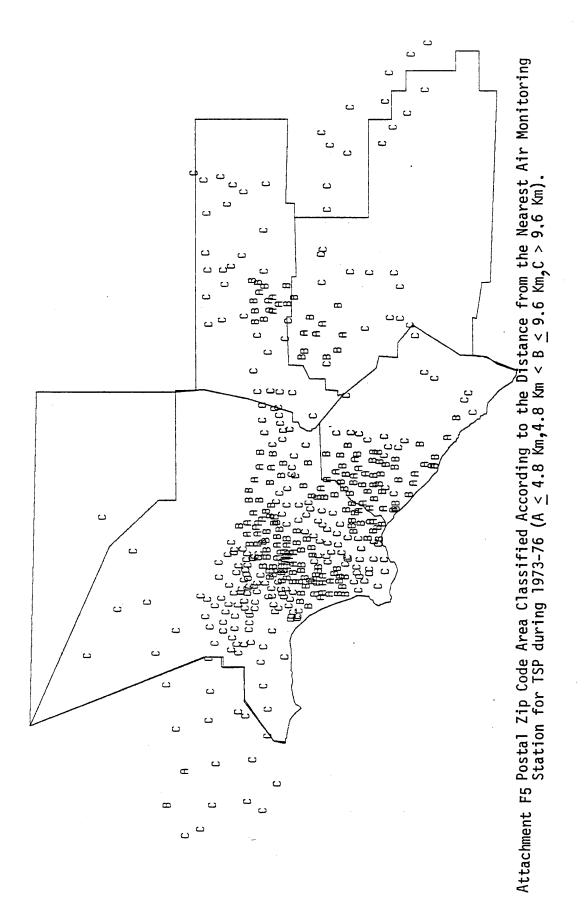
Attachment F2

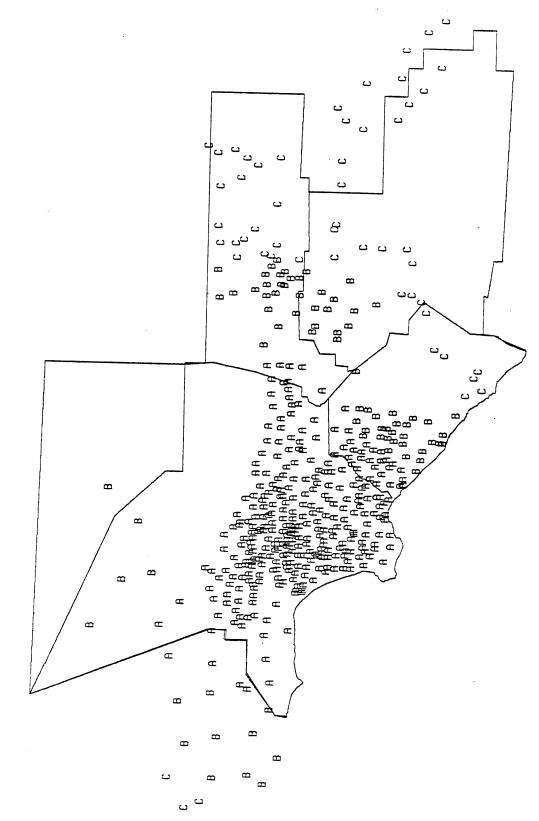


Attachment F3 Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for CO during 1973-76 (A \leq 1.6 Km, 1.6 Km < B \leq 3.2 Km, C > 3.2 Km).



Postal Zip Code Area Classified According to the Distance from the Nearest Air Monitoring Station for $\rm SO_2$ during 1973-76 (A < 4.8 Km,4.8 Km < B < 9.6 Km,C > 9.6 Km). Attachment F4.





Postal Zip Code Area Classified According to the Distance of the Nearest Air Monitoring Station for $S0_4$ during 1973-76 (A \leq 32 Km,32 Km < B \leq 64 Km,C > 64 Km). Attachment F6

\$00 .9106 .8750 .8250 .7750 .7250 .6500 .5500 .4000 .2000 .0000 .9200 .9000 .8500 .8000 .7500 .7000 .6000 .5000 .3000 .1000 MEAN

PERCENTILE VALUES

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	دير	\Z	의	N	MAX	MIN	GMEAN	G.SD	ACTUAL	VALUES					•
			, ,												
	7000	171	1 ?	5	240.0	161-0	190.9	1-2	198.0	240-0	161-0	197.0	168 0		•
	7000	171	1 8	7	206.0	130 0	170.0	1 2	167 0	102 0	107 0	130.0	177.0	407.0	20% 0
			_	Ä	168.0		134.9	1 - 6	107.0	440.0	165.0	130.0			
					181.0							122.C	91.0		
							131.9	1 - 4		142.0		128.0	181.0	134.0	
					304-0		168.7	1.7	179.0	304.0	66.0	200.0	249.0	129.0	
	7000	171	112	5	158.0	29.0	87.2	1.9	96.0	158.0	106-0	108.0	29.0		
	700 G	172	2 1	4	236.0	15510	196.2					155.0	£. / • U		
	2000	172	2	6	302 0	103.0	180.9								
	7000				204.0	4470	10057	1 = 4	102-0	100-0	103.0	302.0	184.0	235.0	
)				201.2	7-5	301-0	163.0	253.0	201-0	181.0	147.0	
	7000					124.0	140.9	1.2	141-0	124-0	125.0	175.0	145.0		
A	7000	172	2 5	5	192.0	74:0	115.8	1.4		192.0		113.0			
A	7000	172	2 6	5	170.0		135.9				122 0	136.0	130 0	-	
					163.0		120.4	1 7	112 0	116.0	122-0			4	4.5
					166.0										163.0
							122.3			166.0			132.0	125.0	
					150.0		103.1	1.3	94.0	62.0	150.0	102.0	158.0	125.0	
					172.0	63.0	92.9	1-4	84.0	63.0	74.0		172.0	118.0	
l M	7000	172	11	6	223.0	70.0	115.8	1.6		135.0	86.0	84.0	_	158.0	
					150.8		110.5		114.0						
							196.3				130-0	120.C	143.0	135.0	
ľ	1000	UII	(2	251-0	137.0	190.5	7-2	193.0	218.0	;137.0	219.0	231.0		
I^:	rugo	UZI	8	1	255.0		194.3	1.2	150.0	222.0	235.0	176.0	152.0	196.0	255.0
					261.0	90.0	148.5	1.4	133.0	261.0	160-0	174.0	90.0		
A.	1006	071	10	6	164.0	92.0	112.8		164.0	92.0		126.0		119.0	
					418.0		190.8	1.9		418.0					
					120.0							215.0		152.0	•
						28.0	71.5	1.7		104.0	74.0	99.0	28.0	52.0	
					342.Ū		207.1	1 _ 4		178.0		234.0	342.0	150.0	
					318.0	84.0	163.3	1.6	134-0	130.0	84.0	318.0	175.0	233-0	•
A7	' 006(072	3	5	248.0	142.0	201.3					180.0	1/2 0	233.0	
5 A 7	0066	072	4	6	180.0	115.0		1 2	155 0	177 0	115 0	127.0	446.0	400 0	
	0060					103.0		4 1	722.0	407.0					
										103.0		184.0	151.0	169.0	
	0000	J (Z	_ O	2	272.0	132.0	198.6	1.3	272.0	132.0	196.0	236.0	186.0		
#[A/	0000	372	7	7	230.0	110.6	178.3	1.3	190.0	226.0	230.0	110.0	140-0	198-0	190 - 0
< 47	ปียิงเ	372	8	6	292.0	107.0	171.5	1-4	195 - 0	292-0	146_0	107.0	171 0	167 0	() 5 5 0
				6	246.0		136.8	1.5	82 0	94 0	244 0	132.0	474 0	300 0	
A7	้อกรถ	772	10	6	208.0		105.0		122-0	100 0		154.0	140.0	400 · U	
12	0060	・・~	11	2	255.0						89.0		208.0		
1,7	0000 0000	J [_	11	·	(J) • U	45.0	92.3		255.0		72.0			110.0	
1	300 C	162	12	6	176.0	36.0	98.6		120.0	36.0	98.0	110.0	176.0	112.0	
[47	J076	571	7	5	150.0	118.0	137.3				118_0	150.0	138 0	• •	
PA7	0076	571	3	7	253_0	114-0	146.9	1_3	114 0	153 0	135 0	125.0	127 0	257 0	111 0
PA7	UGTA	571	9	6	182.0		128-0	1 7	102 D	151 7	447 0	170.0			101.0
PAZ	กกร	(71	10	4	407 0	77 W	120-0	1-3	102.0	151.0	11/.0	158.0	97.0	182.0	
100	00/0	111	IU	0	187.0			1.2	187.0	160.0	108.0	152.0	122.0	158.0	
100	UU / 6	110	77	6	276.0		166.6	1.7	142.0	226.0	66.0	203.0	276.0	180_n	
[A/	UU7 6	71	12	6	199.0	42.0	110.6			199.0		108-0	42.0		
YA Z	0076	572	1	6	303.0	128-0	184.4	1 7	303.0	177 0	105 0	166.0	177 0	430 0	
PA 7	0074	77	· 2	۲.	229.0			1 /	220 0	450.0	イフラーへ	334	111.5		
117	ひひさん	77	7	2	470 0	97.0	170.7	1.4	427 LU	139.0	155.0	221.0	112.0	97.0	
17	UU / 0	276	5	0	179.G	109.0	139.6	1.2	179.0	113.0	159.0	141.0	150.0	109.0	
146	¥″ 16	72	4	6	165.0	137.0	150.7	1.1	146.0	154.0	165_0	143.0	161.0	137.0	
101	ひしてく	:77:	5	6	166 ₋ 0	97.0	137.9	1.2	129.0	97 0	154 0	138.0	166.0	156 D	
PA7	0076	372	. 6	6	162.0	137 n	145 6	1 1	1/1 0	1// 0	157 0	120 att	100.0	170.0	
PAZ	0074	77	7	7	143.0	74 0	100 /	4 7	407 0	144-0	122-11	137.0			
PA >	ひひつく	72	(7	143.0	71.0	108-0			111.0		/1.C	102.0	143.0	142.0
1	UU 7 6	12	×	Ó	173.0	192.0	132.5	1.2	102.0	171.0	121.0	173.0	115.0	129.0	
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Pollu Metho Stati Code Year Month	N MAX	MIN	GMEAN	6-20	ACTUAL	VALUES	•			
\$04A7000173 1 J		4.5	7.8	1 - 4	8.3	6.1	10_8	8.0	11.3	4.5
	6 10.0	3.1	5.1	1.5	4.4	4.3	4.9	6.3	3.1	10.0
S04A7G00173 3		5.1	7.0	1 - 4	7.8	6.2	5.9	12.2	6.4	5.1
	6 50.3	2,-4	13.0	3.7	2.4	21.2	37.5	18.7	2.7	50.3
	3 72.2	6.1	18.2	2.7	6.1	9.7	42.0	72.2	23.2	8.7
	52.0	5.2	12.7	2.2	5.2	16.7	11.4	7.7	10.8	52.0
S04A7G00173 7		16.7	29.4	1.4	42.2	35.7	42.6	23.4	16.7	24.5
S04A7000173 8 6		11.9	17.9	1.5	30.6	19.6	25.8	12.0	11.9	14.8
S0447000173 9 6		6-1	18.0	2.0	13.9	14.2	42.8	36.1	6.1	18.3
\$04A700017310 8		4.0	7.1	1.9	22.5	4.0	6.5	6.5	4.0	8.6
S04A700017311 (0.1	2.4	9.8	3.6	39.0	23.2	0.1	0.5	1.3
S04A700017312		. •Ci • 7	4.9	3.5	3.3	0.7	16.9	11.1	2.3	13.1
\$04A7000174 1 6		1.9	3.9	1.7	1.9	2.7	3.9	\$. 8	4.3	4.5
\$04A7000174 2 6		3.2	5.6	2.2	24.2	4.2	3.6	5.3	3.2	3.3
S04A7000174 3 8		1.8	4.7	2.4	1.9	1.8	8.6	12.9	8.6	3.1
\$04A7900174 4 6		.2.6	6.7	2.0	2.6	8.8	5.0	13.9	16.2	4.5
\$04A7000174 5 6	_	4 - 1	12.7	2.3	16.7	33.8	25.6	4.1	5.1	13.7
504A7000174 6 5		3.9	15.6	3.1	21.2	42.3	44.5	3.9	5.9	
\$04A70U0174 7 6		7.7	12.6	1.6	15.4	20.0	8.9	8.8	7.7	21.9
S04A7G00174 8 6		9.4	18.6	1.6	9.4	11.1	23.1	28.6	27.0	22.0
S04A7500174 9 6		16.2	24.8	1-4	16.2	32.4	16.8	38.9	23.1	29.1
\$04A700017410 6		2.7	11.6	2.5	12.3	8.5	38.4	23.5	9.5	2.7
'S04A700017411 6		1.5	3 - 4	3.0	2.6	4.1	1.5	27.4	1.6	2.3
S04A700017412 6		2.5	4.4	1.9	5.7	4.4	13.3	3.3	2.5	2.5
S04A7000175 1 6		2.3	3.5	1.5	3.2	5.7	2.3	2.3	6.0	3.0
S04A7000175 2 4		2.9	4.2	1.5	2.9	4.6	3.2	7.4		
S04A7000175 3 6		3.0	7.1	2-4	36.3	3.0	3.9	7.7	4.8	8.3
S04A7000175 4 5		2.5	6.7	1.9	.:2.5	6.0	6.8	8.4	15.6	
S04A7G00175 5 5		4.3	12.3	2.4	4.3	26.8	5.5	14.7	30.4	
\$04A7000175 6 5		13.6	18.7	1.3	13.6	23.9	14.4	23.5	20.9	
\$04A7300175 7 5		12.2	14.8	1.3	13.7	15.5	12.2	22.4	12.3	
S04A7000175 8 5		8.2	15.4	1.5	20.4	22.8	17.4	13.2	8.2	
S04A7000175 9 5	-	11.3	17.8	1.6	11.9	16.1	31.0	11.3	26.8	
\$04A700017510 5		3.4	8.7	1.9	13.4	7.5	3.4	19.7	7.5	
\$04A700017511 5		3.5	9.5	2.3	21.0	22.5	8.7	3.5	5.5	
\$04A700017512 5		5.7	8.9	1.8	5.8	5.9	5.7	20.9	14.0	
\$04A7000176 1 6	23.5	5-1	11.0	1.7	5.1	9.6	9.8	23.5	16.0	10.0
\$04A7000176 2 4	15.0	5.8	7.6	1-6	5.8	6.1	15.0	6.4		
\$0447000176 3 6	15.8	4.6		1.6	5.6	5.6	4.6	8.0	15.8	6.5
\$04A7000176 4 5	17.0	5.4	8.7	1.6	5.4	8.8	7.6	17.0	10.5	
\$04A7000176 5 5		9.6	13.2	1.3	9.6	18.3	14.9	12.8	11.9	
\$04A7UU0176 6 5 \$04A7UU0176 7 4	14.3	8.2		1.2	11.5	8.2	13.4	10.3	14.3	
	34.4	16.5	23.4	1-4	34.4	19.9	16.5	26.4		
	29.6	7.8	13.1	1.7		17.4	8.3	7.8	29.6	
	23.8	7.7	15.8	1-6	18.8	23.2	12.4	23.8	7.7	
	28.8	4.2	10.0	2.2	7.8	4.2	28.8	10.6		
	16.2	4.3	9.3	1.7	4.3	8-0	8.8	16.2	13.9	
[4A700017612 5	13.2	2.4	6.0	1.9	2.4	6.7	8.0	13.2	4 . 4	
\$54A7007673 1 6		4.7	7.1	1.5	6.5	5.5	14.0	7.3	7.2	4.7
\$04A7007673 2 6	11.9	3.4	6.3	1.5	6.8	5.7	5.5	7.3	3.4	11.9

Attachment I Sulfate Air Quality Data Base

	MEAN	DOSE	F>10	E>10	F>15	E>15	F>20	E>20	F>25	E>25
OXIA7000173 1 1 OXIA7000173 2 1 OXIA7000173 3 1 OXIA7000173 4 1 OXIA7000173 5 1 OXIA7000173 6 1 OXIA7000173 7 1 OXIA7000173 9 1 OXIA7000173 9 1 OXIA700017310 1 OXIA700017311 1 OXIA7000174 1 1 OXIA7000174 2 1 OXIA7000174 2 1 OXIA7000174 5 1 OXIA7000174 6 1 OXIA7000174 7 1 OXIA7000174 8 1 OXIA7000174 9 1 OXIA7000174 9 1 OXIA7000174 1 1 OXIA7000174 1 1 OXIA7000175 1 1 OXIA7000175 1 1 OXIA7000175 2 1 OXIA7000175 6 1 OXIA7000175 7 1 OXIA7000175 8 1 OXIA7000175 9 1 OXIA7000175 1 1 OXIA7000175 2 1 OXIA7000175 1 1 OXIA7000175 1 1 OXIA7000175 2 1 OXIA7000175 3 1 OXIA7000175 3 1 OXIA7000175 1 1 OXIA7000175 1 1 OXIA7000175 2 1 OXIA7000175 3 1	111234440668931133075058097084557626813407760032139	967 940 1264 1728 3168 2578 2578 2578 2578 2578 2578 2578 257	0 0 0 3 7 9 7 4 2 7 9 2 4 2 9 7 9 3 6 4 3 0 1 6 4 7 7 7 6 1 1 8 9 3 7 0 1 6 3 4 7 4 3 0 1 6 4 7 7 7 6 1 1 3 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 4 3 0 1 6 3 4 7 6 6 4 7 6 6 4 7 7 6 6 7 7 7 6 6 7 7 7 6 7 7 7 6 7	0 0 0 1 7 9 3 8 4 7 7 5 0 0 0 6 4 1 1 9 1 6 1 8 1 8 1 9 1 9 1 9 1 9 1 9 1 9 1 9 1 9	00001220536300076748205500101012222640003321129270	0000029035890008961838600001004859600003225116160 731635811000001004859600003225116160	00005930012000001534700000000237100000037134300	00006750015000002135900000005091000000067596000 217	9990247999999999999999999999999999999999	00005220000000000000000000000000000000

Attachment J. Oxidant Exposure Parameter Data Base.

	MEAN	DOSE	F>5	E>5	F>15	E>15	F>20	E>20	F>25	E>25
NO2A700017412 1 NO2A7000175 1 1 NO2A7000175 2 1 NO2A7000175 3 1 NO2A7000175 4 1 NO2A7000175 5 1 NO2A7000175 6 1 NO2A7000175 7 1 NO2A7000175 8 1 NO2A7000175 9 1 NO2A7000175 9 1 NO2A700017510 1 NO2A700017511 1 NO2A7000176 1 1 NO2A7000176 2 1 NO2A7000176 3 1 NO2A7000176 5 1 NO2A7000176 6 1 NO2A7000176 7 1 NO2A7000176 8 1 NO2A7000176 9 1 NO2A7000176 9 1 NO2A7000176 1 1 NO2A7000176 9 1 NO2A7000176 1 1	9.2 10.5	6032 6032	66937600446603255555555555555555555555555555555555	2496 1039 540 1017 1740 1206 1241 1377 3414 1683 2588 2420 1458 1600 1272 2249 4842 4922 4922 4922 4922 4922 492	376 0 123 144 145 145 145 147 142 145 147 142 145 147 142 145 147 142 145 147 142 145 147 142 145 145 147 142 145 145 147 142 145 145 145 145 145 145 145 145 145 145	215 10 40 83 126 57 763 97 77 763 97 77 763 97 77 763 97 77 763 97 77 763 97 763 97 763 97 763 97 763 97 763 97 97 97 97 97 97 97 97 97 97 97 97 97	14002604638758305377778903383102232981565331281902 11565331281902	104 0 0 3 44 17 0 2 3 9 6 6 5 0 6 6 5 0 19 1 8 4 0 4 5 2 10 6 4 7 6 8 3 9 1 8 2 2 6 3 6 5 1 8 2 1 5 1 8 2 2 6 3 6 5 1 8 2 1 8 4 0 4 5 2 1 8 6 7 4 7 6 8 3 9 1 8 2 8 6 7 6 7 6 8 7 8 8 7 8 8 8 7 8 8 8 8 7 8 8 8 8	9000330202050002330028840000029342040023090491	4600012808050000587082225000000000000000000000

CO1A7000173 1 1 CO1A7000173 2 1 CO1A7000173 3 1	MEAN 5.6 4.9 3.7	DOSE 4166 3292 2752	F>5 372 302 186	E>5 1096 672 249	F>10 74 40 7	E>10 259 83 5	F>15 14 3 0	E>15 54 5 0	F>20 5 0	E>20 13 0 0
C01A7000173 4 1 C01A7000173 5 1 C01A7000173 6 1 C01A7000173 7 1 C01A7000173 8 1	4.1 4.0 4.6 4.0 4.2	2952 2976 3312 2976 3124	216 223 288 204 223	338 276 505 226 273	14 7 28 7 11	18 14 66 8 16	0 0 5 0	0 9 0	0 0 0	0
C01A7000173 9 1 C01A700017310 1 C01A700017311 1 C01A700017312 1	4.6 6.5 5.6 7.1	3312 4836 4032 5282	252 372 360 520	539 1748 1101 1877	36 148 72 148	96 539 210 501	0 5 44 14 37	0 3 146 19 92	0 0 11 0 3	0 0 37 0 5
CO1A7000174 1 1 CO1A7000174 2 1 CO1A7000174 3 1 CO1A7000174 4 1 CO1A7000174 5 1	6.4 4.9 3.3 2.9 2.4	4761 3292 2455 2088 1785	446 268 130 108 59	1661 793 303 108 63	130 53 22 3	378 157 44 2 0	29 13 3 0	90 26 6 0	3 0 0 0	3 0 0 0
C01A7000174 6 1 C01A7000174 7 1 C01A7000174 8 1 C01A7000174 9 1 C01A700017410 1	2.9 3.8 3.4 5.6 5.5	2088 2827 2529 4032 4092	108 204 167 360 372	152 261 232 968 1039	7 7 7 64 67	18 6 3 171 253	1 0 0 14 22	0 0 0 30 85	0 0 0 1 7	0 0 0 3 22
C01A700017411 1 C01A700017412 1 C01A7000175 1 1 C01A7000175 2 1 C01A7000175 3 1	7.5 8.5 8.4 6.0 4.5	5400 6324 6249 4032 3348	504 520 520 336 260	2193 2919 2863 1258 653	180 223 223 100 44	690 1115 1096 342	57 93 93 23	179 361 380 80	10 29 29 6	44 77 123 18
C01A7000175 4 1 C01A7000175 5 1 C01A7000175 6 1 C01A7000175 7 1	3.2 3.2 3.3 2.6	2304 2380 2376 1934	144 148 162 74	270 195 243 82	12 7 10 3	97 39 14 34 8	7 4 0 3 0	19 4 0 6 0	0 0 0 0	0 0 0 0
C01A7000175 8 1 C01A7000175 9 1 C01A700017510 1 C01A700017511 1 C01A700017512 1	3.1 4.1 4.4 5.9 7.0	2306 2952 3273 4248 5208	111 252 260 360 446	197 462 644 1428 2043	7 28 37 108 186	16 102 121 429 564	0 9 9 36 52	0 24 26 121 99	0 1 1 10 3	0 0 2 35 10
C01A7000176 1 1 C01A7000176 2 1 C01A7000176 3 1 C01A7000176 4 1 C01A7000176 5 1	6.7 4.0 3.8 3.3 2.5	4984 2688 2827 2376 1860	446 201 223 144 74	1743 509 350 173 93	148 26 14 5 0	423 78 19 10	29 5 0 0	73 11 0 0 0	5 0 0 0	7 0 0 0
C01A7000176 6 1 C01A7000176 7 1 C01A7000176 8 1 C01A7000176 9 1 C01A700017610 1	3.2 3.0 3.3 3.2 4.3	2304 2232 2455 2304 3199	126 74 148 144 260	180 74 349 221 564	10 1 22 7 29	30 0 76 14	1 0 4 0	3 0 16 0	0 0 1 0	0 0 0 0
CO1A700017610 1 CO1A700017611 1 CO1A700017612 1	6.2 6.7	4464 4984	360 446	1477 1743	126 148	56 352 423	2 28 22	7 52 62	0 4 1	0 7 0

	MEAN	DOSE	F>2	E>2	F>4	E>4	F>8	E>8	F>14	E>14
S02A7000176 1 1 S02A7000176 2 1 S02A7000176 3 1 S02A7000176 4 1 S02A7000176 5 1 S02A7000176 6 1 S02A7000176 7 1 S02A7000176 8 1 S02A7000176 9 2 S02A700017610 1 S02A700017611 1	1.6 2.4 2.1 1.3 1.4 1.7 1.2 2.7 3.5 7 4.9 1.4 6.3 0.0 0.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0 1.0	1440 1488 1656	262 302 163 163 163 163 163 163 163 163 163 163	958 813 813 813 814 814 815 814 815 815 816 816 817 817 817 817 817 817 817 817	763141693997621151729372780007219344924420354811402993 111693997621151729372780007219344924703548114002035481175020000000000000000000000000000000000	020206998459846000022263922124042808669044120232341 1568845984000022263922124042808669044120232341	00005111017000000000001350000000000102001000010	0000M0000M00000000000M0000000000000000	000000000000000000000000000000000000000	000000000000000000000000000000000000000

Attachment M. SO₂ Exposure Parameter Data Base

Attachment N. TSP Exposure Parameter Data Base

Attachment O. Sulfate Exposure Parameter Data Base

Attachment P

Cubic Spline Interpolation Scheme With A Parabolic Leapfrog Correction

Reference

- Sommers W.T. (1976), "Data Interpolation by Cubic Splines", USDA Forest Service Research Note PSW - 313
- Guardado, J.L. and W.T. Sommers (1977), "Interpolation of Unevenly Spaced Data Using a Parabolic Leapfrog Correction Method and Cubic Splines", USDA Forest Service Research Note PSW - 324.

PACIFIC SOUTHWEST Forest and Range Experiment Station

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DATA INTERPOLATION BY CUBIC SPLINES

William T. Sommers

USDA Forest Service Research Note PSW-313 1976

Sommers, William T.

1976. Data interpolation by cubic splines. USDA Forest Serv. Res. Note PSW-313, 5 p., illus. Pacific Southwest Forest and Range Exp. Stn., Berkeley, Calif.

The cubic spline is a cubic polynomial that can be used with data sets of either fixed or variable mesh length. It is useful for interpolating data, smoothing data, estimating volume, and achieving greater precision in numerical analysis. The technique has a variety of applications in forestry and related disciplines. A FORTRAN IV computer program, available upon request, will yield a bi-cubic spline interpolation for either a variable mesh length or an equal interval grid.

Oxford: 0-015.5

Retrieval Terms: data smoothing; forestry; fire management; cubic splines.

Cubic splines are a mathematical formulation of a mechanical technique used by draftsmen for many years, in which they fit a flexible strip of material to known points in order to form a smooth curve. For mathematical purposes, the known points are data values on a fixed or variable mesh and the flexible strip is approximated by a piecewise continuous function. This function is defined as a cubic polynomial in each interval between the known points, with continuous first and second derivations.

The cubic spline is useful as a data interpolation technique, as a finite element method in numerical analysis, as a data smoother and as a volume estimator. Meteorologists have used cubic splines for data smoothing¹, for telescoping a grid², and for evaluating derivatives in numerical analysis computations.³ Cubic splines are of potential use in forestry and fire applications that deal with numeric data bases. Examples of such use include smoothing or interpolating digital topography, working with brush cover heights and classes, estimating volume contained between tree tops and ground level from sparse data, interpolating growth rates between known points, and in other applications dealing with a spatial or time series data base.

SINGLE CUBIC SPLINE

The basic mathematical theory for both the single cubic spline (one-dimensional) and bi-cubic spline (two-dimensional) is explained by Ahlberg and others.⁴ A concise reformulation of the basic theory tailored to application follows.

If $S_d(f;x)$ is the spline of interpolation of the function f(x) on the mesh d, then

$$\int_{1}^{b} |f''(x)|^{2} dx =$$

$$\int_{a}^{b} \left| s_{d}^{"}(f;x) \right|^{2} dx + \int_{a}^{b} \left| f''(x) - s_{d}^{"}(f;x) \right|^{2} dx \qquad (1)$$

where primes represent differentiation. The integral on the left hand side of (1) is minimized by the spline fit and end conditions s_d^* (f;a) = s_d^* (f;b)= 0. In words, the cubic spline represents the curve with minimum curvature that will fit the given data, be continuous through the second derivative and assure the "smoothest" fit to the data.

The derivations which follow may be found in expanded form in the publication by Ahlberg and others $\frac{4}{}$. Consider the mesh d of length b-a subdivided such that

d: $a = x_0 < x_i < ... < x_n = b$ where the spacing between the x_i need not be uniform. Data values at the x_i points are given by

$$Z: z_0, z_1, \ldots, z_n$$

and may represent digital elevation values or some other form of data on a grid. We seek a spline function $S_d(x)$ which is continuous through its second derivative $(S_d^u(x))$, coincides with a cubic in each subinterval $x_{i-1} \le x \le x_i$ ($i=1,2,\ldots,n$) and satisfies $S_d(x_i) = z_i$ at the given points x_i . $S_d(x)$ is periodic or nonperiodic depending on whether the condition $S_d(a+) = S_d(b-)$ is satisfied or not.

We designate \mathbf{M}_i the moment \mathbf{S}_d^* and from the prescribed linearity of the second derivative we write

$$S''_d(x) = M_{i-1}\left(\frac{x_i-x}{h_i}\right) + M_i\left(\frac{x-x_{i-1}}{h_i}\right)$$
 (2)

where $h_i = x_i - x_{i-1}$ is the spacing between the known data points. Equation 2 simply restates the minimum curvature property and uses the prescribed linear change of the second derivative between known points x_i .

Integrating (2) twice and evaluating the constants yields

$$S'_{d}(x) = -M_{i-1} \left(\frac{(x_{i}-x)^{2}}{2h_{i}} \right) + M_{i} \left(\frac{x-x_{i-1}}{2h_{i}} \right)^{2} + M_{i} \left(\frac{x-x_{i-1}}{2h_{i}} \right)^{2}$$

$$\left(\frac{z_{i}-z_{i-1}}{h_{i}}\right)\left(\frac{M_{i}-M_{i-1}}{6}\right)h_{i} \tag{3}$$

$$S_{d}(x) = M_{i-1} \left(\frac{(x_{i}-x)^{3}}{6h_{i}} \right) + M_{i} \left(\frac{(x-x_{i-1})^{3}}{6h_{i}} \right) + \left(\frac{x_{i-1}}{6h_{i}} - \left(\frac{M_{i-1}}{6} h_{i}^{2} \right) \right) \left(\frac{x_{i}-x}{h_{i}} \right) + \left(\frac{x_{i}}{6} - \left(\frac{M_{i}h_{i}^{2}}{6} \right) \right) \left(\frac{x_{i}-x_{i-1}}{h_{i}} \right)$$

$$(4)$$

Equations 3 give the slope of the spline and equations 4 the spline itself. We will continue the development here in terms of the moments, M_i , but one could just as easily develop the technique in terms of the slopes. One sided limits of (3) wield

$$S_{d}^{i}(x_{i}^{-}) = \left(\frac{h_{i}}{6} M_{i-1}\right) + \left(\frac{h_{i}}{3} M_{i}\right) + \left(\frac{z_{i}^{-z_{i-1}}}{h_{i}}\right)$$

$$S_{d}^{i}(x_{i}^{+}) = -\left(\frac{h_{i+1}}{3} M_{i}\right) - \left(\frac{h_{i+1}}{6} M_{i+1}\right) + \left(\frac{z_{i}^{-z_{i-1}}}{h_{i+1}^{-z_{i}}}\right)$$

$$(5a)$$

These and the slope continuity condition yield

$$\left(\frac{h_{i}}{6} M_{i-1}\right) + \left(\frac{h_{i} + h_{i+1}}{3} M_{i}\right) + \left(\frac{h_{i+1}}{6} M_{i+1}\right) =
\left(\frac{z_{i-1} - z_{i}}{h_{i+1}}\right) - \left(\frac{z_{i} - z_{i-1}}{h_{i}}\right)$$
(6)

which, for the periodic spline, give N-1 simultaneous equations for the moments M_1 , M_2 ,.... M_n . For the nonperiodic spline, additional, or end, conditions must be supplied. These conditions are a matter of experiment. We have found that specifying the slope at the end points (a,b) by linearly extrapolating the z values from (z_1, z_2) and (z_{n-1}, z_n) works well. Generalized end conditions are

$$2M_0 + \lambda_0 M_1 = C_0 \tag{7a}$$

$$\mu_{n}M_{n-1} + 2M_{n} = C_{n} \tag{7b}$$

in which

$$\lambda_{\hat{1}} = \frac{h_{\hat{1}+\hat{1}}}{h_{\hat{1}}+h_{\hat{1}+\hat{1}}} \tag{8a}$$

$$\mu_{i} = 1 - \lambda_{i} \quad (i = 1, 2,, n-1)$$
 (8b)

Equations 6 now become

$$\mu_{i}^{M}_{i-1} + 2M_{i} + \lambda_{i}^{M}_{i+1} = 6$$

$$\left[(z_{i+1} - z_1)/h_{i+1} \right] - \left[(z_i - z_{i-1})/h_i \right]$$
(9)

Writing equations 7 and 8 in matrix form as

$$\begin{vmatrix} \overline{2} & \lambda_{0} & 0 & \cdots & 0 & 0 & 0 & 0 \\ \mu_{1} & 2 & \lambda_{1} & \cdots & 0 & 0 & 0 & 0 \\ 0 & \mu_{2} & 2 & \cdots & 0 & 0 & 0 & 0 \\ \vdots & \vdots & \vdots & \vdots & \vdots & \vdots & \vdots \\ 0 & 0 & 0 & \cdots & 2 & \lambda_{n-2} & 0 & M_{n-2} \\ 0 & 0 & 0 & \cdots & \mu_{n-1} & 2 & \lambda_{n-1} \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & \mu_{n} & 2 & M_{n} & \vdots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & 0 & \dots \\ 0 & 0 & 0 & \cdots & 0 & \dots \\ 0$$

or
$$AM = C$$
 (10b)

shows them to be a tri-diagonal system of equations which can be solved by numerical techniques. Repeated calculations on the same mesh (i.e., only the z_i change) can yield considerable time savings due to the non-variance of h_i , λ_i , and μ_i .

Use of equal interval meshes can save even more computational time. In many meteorological applications, large iterative computations are carried out on equal interval, multi-level grids. As a result much time can be saved. For an equal interval mesh, the inverse of the coefficient matrix A of equations 10 takes on simplified form. Equation 10b may be rewritten as $M = A^{-1}C$, and the moments then may be calculated by inverting the coefficient matrix A. The nxn determinant of A in terms of constant λ is

$$D_{n} (\lambda) \begin{vmatrix} 2 & \lambda & 0 \dots & & & & \\ 1-\lambda & 2 & \lambda & & & & & \\ 0 & \lambda & 2 & & & & & \\ & & \dots & & 2 & 0 & 0 & \\ & & & 1-\lambda & 2 & \lambda & \\ & & & 0 & 1-\lambda & 2 & \\ \end{vmatrix}$$
(11)

since $\lambda = \frac{h_i}{h_i + h_{i+1}} = 1/2$ for equal intervals, the

determinant $D_n = D_n$ (1/2) yields

$$D_{n} = \frac{\left(\frac{1+3}{2}^{1/2}\right)^{n+1} - \left(\frac{1-3}{2}^{1/2}\right)^{n+1}}{3^{1/2}}$$

$$D_{0} = 1$$

$$D_{-1} = 0$$
(12)

The coefficient determinant |A| satisfies

$$|A| = 4D_{n-1} - 1/2D_{n-2}$$
 (13)

And the elements of the inverse of the coefficient $\ensuremath{\mathsf{mat}}_r$ are

$$A_{i,j}^{-1} = \frac{(-1)^{i+j} \left(2D_{j-1} - 1/4D_{i-2}\right) \left(2D_{n-j-1} - 1/4D_{n-j-2}\right)}{2^{i-j} |A|}$$

$$(0 < j \le i \le N)$$
 (14)

$$A_{i,j}^{-1} = \frac{(1)^{i+j} \left(2D_{j-1} - 1/4D_{j-2}\right) \left(2D_{n-i-1} - 1/4D_{n-i-2}\right)}{2^{i-j} |A|}$$

$$A_{0,j}^{-1} = \frac{(-1)^{j} \frac{1}{2} \left(2D_{n-j-1} - \frac{1}{4D_{n-j-2}}\right)}{2^{j-1} |A|}$$
 (14b)

$$A_{N,j}^{-1} = \frac{(-1)^{N+j} \frac{1}{2} \frac{2D_{j-1} - 1/4D_{j-2}}{2^{N-j-1} |A|}$$
 (0 < j \le N)

$$(0 < j < N)$$
 (14d)

Since
$$M_i = \sum_{j=0}^{N} A_{i,j}^{-1} C_j$$

we get
$$M_i = A_{i,o}^{-1} \left[C_o + \frac{6}{h^2} \left(Z_o + \lambda_o Z_l \right) \right] - \frac{18}{h^2} \sum_{i=0}^{N} A_{i,j}^{-1} Z_j$$

$$+\frac{6}{h^2}Z_i + A_{i,N}^{-1} \left[C_N + \frac{6}{h^2} \left(Z_N - \mu_N Z_{N-1} \right) \right]$$
 (16)

Additional shortcuts can be used when the spline is use operationally on an equal interval mesh when the Z fiel does not vary drastically from run to run. Since the $A_{i,j}^{-1}$ decay rapidly with distance from the principal diagonal (i.e., abs(i-j) becomes large), experimentation should yield suitable truncation methods.

Bi-cubic Spline

We have dealt so far with the single cubic spline. For a bi-cubic spline, extension to a grid of k rows by columns is easily accomplished by considering the calculated $S_d(x)$ values to be new z values and rotating the single cubic spline technique through 90° . For example, consider that you wish to interpolate to a point κ, γ between rows k and k+l and columns l and l+l. Calculate $S_d(\gamma)$ for all rows and calculate $S_d(\kappa, \gamma)$ for the interpolated values $S_d(\gamma)$ in the " γ column".

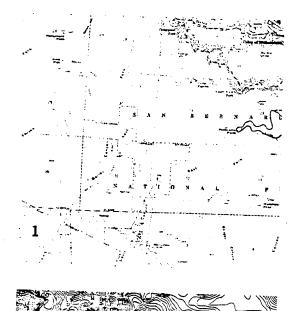
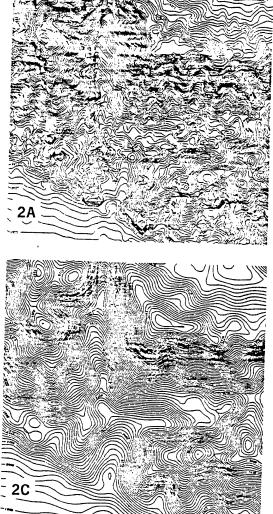


Figure 1-Section of USGS San Bernardino north quadrangle 7.5-minute topographic map shows the Devil Canyon area of the San Bernardino National Forest, in southern California.



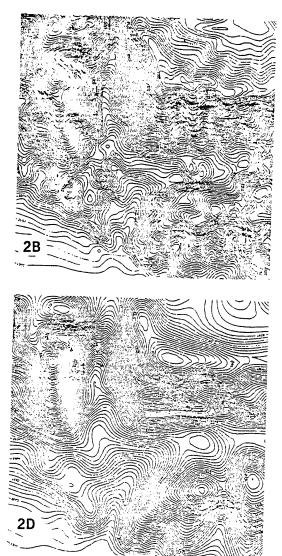


Figure 2—Computer generated plots of digitized topography for the area shown in Figure 1. The plots are based on elevation values at gridpoints on a 400-foot (122-m) mesh. The gridpoint data represents actual values (2A), values interpolated from an 800-foot mesh (2B), values interpolated from a 1200-foot mesh (2C), and values interpolated from a 1600-foot mesh (2D).

APPLICATION

The cubic spline may be used to interpolate data or to smooth data, to estimate volume or to replace or supplement finite difference approximations in numerical calculations. Before applying the spline technique for interpolation, the user must understand the nature of the available data and remember that the spline is a technique for achieving a polynomial fit with minimum curvature. If the actual data contain significant variations, or frequencies (wavelengths) that are higher than sample data collection, then the root mean square error associated with spline interpolation is not likely to be significantly better than the root mean square error associated with a linear interpolation scheme. The spline interpolation scheme will provide a minimum curvature fit, to the data, that is continuous through the second derivative. Spline interpolation is advisable, therefore, when the interpolated data are to be used in calculations which may be adversely affected by discontinuous slopes.

As an example of data interpolation and smoothing, we applied the bi-cubic spline technique to digital terrain heights on a grid. The topography of the Devil Canyon area of the San Bernardino National Forest in southern California (fig. 1) had been hand-digitized at a horizontal spacing of 400 feet (122 m) and an elevation resolution of 20 feet (6 m). A meteorological research network is located there and is used in the development of numerical weather models which also require topographic data. We created new data sets by using every other point (800-foot spacing) of the original grid, every third point (1200-foot spacing), and every fourth point (1600-foot spacing). The new data sets were then interpolated by bi-cubic splines to the original grid points. The root mean square errors in elevation at the interpolated points that resulted were 67 feet at the 800-foot mesh length, 93 feet at 1200 and 119 feet at 1600. Figures 2a, b, c, and d show computer-generated plots of the original 400-foot digitization and the three interpolations. Note how the many small ravines of 500-foot to 1000-foot width that are seen in the quadrangle map (fig. 1) are represented and smoothed in the four plotted digital

maps (fig. 2). If it is assumed that a grid of mesh length Δ can resolve features with a 2 Δ -characteristic scale, then one can expect features of 800, 1600, 2400, 3200-foot widths to be resolved in figs. 2a, b, c, and d. The resolutions we achieved were actually somewhat better. Major topographic features such as the east and west forks of Devil Canyon, the Pine Flat and Camp Paivika plateaus, and Monument Peak are discernible through the 1600-foot mesh (fig. 2d).

Users should determine the balance between resolution and smoothing required for their particular needs and choose an appropriate interpolation interval. If variable resolution is required, the spline can telescope to higher resolution in areas of interest while maintaining lower resolution elsewhere. This technique can be applied in digitizing the data by choosing a tight digitizing interval where gradients are large and allowing looser intervals elsewhere, and then having the variable mesh length spline generate an equal mesh length grid.

A FORTRAN IV computer program, written by Julio L. Guardado, University of California, Riverside, will yield a bi-cubic spline interpolation for either a variable mesh length or an equal interval grid. Copy of the program is available on request to: Director, Pacific Southwest Forest and Range Experiment Station, P.O. Box 245, Berkeley, California 94701, Attention: Computer Services Librarian.

NOTES

- 1 Fritsch, J. M. 1971. Objective analysis of a two-dimensional data field by the cubic spline technique. Month. Weather Rev. 99:379-386.
- ² Price, G. V., and A. K. MacPherson 1973. A numerical weather forecasting method using cubic splines on a variable mesh. J. Appl. Meteorol. 12:1102-1113.
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Interpolation of unevenly spaced data using a parabolic leapfrog correction method and cubic splines

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The technique proposed allows interpolation of data recorded at unevenly spaced sites to a regular grid or to other sites. Known data are interpolated to an initial guess field grid of unevenly spaced rows and columns by a simple distance weighting procedure. The initial guess field is then adjusted by using a parabolic leapfrog correction and the known data. The final output is then generated by use of cubic spline interpolation on the adjusted grid. Application to test and actual data showed acceptably low root-mean-square error.

Oxford: 111.0-015.5.

Retrieval Terms: meteorological analysis; data processing; cubic splines.

Data collection sites used during field experiments or in operational observing networks are usually not located at points of an orthogonal grid. But numerical analysis techniques for processing the collected observations usually require data on an orthogonal grid. Even when data from unevenly spaced points are interpolated to other unevenly spaced points, many techniques require an intermediate orthogonal grid.

In this note, we describe a technique by which data recorded at unevenly spaced sites are used with cubic splines¹ to interpolate to unevenly spaced sites or to generate an orthogonal grid. The technique may be applied to any type of data. Judicious choice of grid spacing will, however, improve performance when the general variation of the specific data being interpolated is known.

PROCEDURE FOR INTERPOLATING

This procedure for interpolating from unevenly spaced data has three steps. First, we use the known data to generate an initial "guess field" grid by means of a simple square-of-distance weighting scheme using straight-line distance between points. Then, we adjust the initial guess field rows and columns using the known data points and a parabolic leapfrogging technique. Finally, we interpolate the corrected field to a desired orthogonal grid or to another set of unevenly spaced points.

Initial guess field rows and columns are chosen such that each row and column passes through one of the known data points. Normally, only one column is generated per unique data point x value and only one row is generated per unique data point y value. The grid produced is orthogonal but has unevenly spaced rows and columns. After the initial rows and columns are adjusted, cubic spline interpolation is performed

independently on each row and column to generate two data values at the desired output points. These two data values are then averaged to produce the final output.

Inverse Square-of-Distance Weighting

The problem of determining a function value at a random point on a grid was solved using an inverse square-of-distance weighting applied to the known point function values. The basic procedure for finding a value at an unknown point from weighted values at n known points is

$$f' = \frac{\sum_{i=1}^{n} w_i f_i}{\sum_{i=1}^{n} w_i}$$

where f' is the function value at the unknown point, f_i are the function values at the n known points, and w_i are the weights given to the f_i . Because we desire that the contributions of the f_i to f' be weighted as the inverse square of the distance, then

$$w_i = \frac{1}{d_i^2}$$

where d_i are the distances between known points and unknown point. This procedure is repeated for each unknown point on the row or column of the initial guess grid which we are filling.

For example, say that the point x_1 was a distance d_1 from a point x' at which we would like to approximate the function value. Furthermore, the point x_2 is twice as far from x' as x_1 . Point x_1 would be given a weight of 4/5 and point x_2 would receive a weight of 1/5 because it is twice as far away as x_1 and should be given 1/4 the weight of x_1 . The sum of the weights multiplied by the function values at the known points yields the function value at the unknown point.

The Parabolic Leapfrog

Since all the initial guess rows and columns pass through at least one known data point, the question "How can we correct the initial guess given a known point?" arises. It was decided to correct the initial rows and columns by a "leapfrog" technique beginning at the known data point and performing piecewise corrections using a parabola passed through three points to obtain a fourth point within that interval. The leapfrogging procedure is diagrammed for a case where (a) leftmost endpoint is known, (b) rightmost endpoint is known, and (c) a "central" point is known. The points marked with an asterisk

are used to determine those marked with a question mark. The numbers signify the leapfrogging sequence.

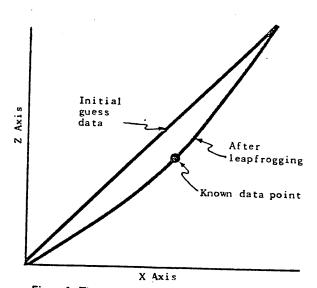


Figure 1—The curve demonstrates the effect of parabolic leapfrogging applied to a straight line and a known data point not on that line.

Figure 1 shows the effect of the leapfrogging procedure on a straight line, given a known point not on the line. We can see by the distance between the two lines that the effect of the correction procedure diminishes as the distance from the known point increases.

Given a set of points

$$(x_i,f_i)$$
 for $i = 0,1,2,3,...n$ $n \ge 3$

and given a known point (x_k, f_k) $0 \le k \le n$, then the remaining f_i $(i = 0, ..., n i \ne k)$ may be corrected by using the technique of parabolic leapfrogging.

Let j be the starting subscript for choosing the points to be used for interpolation and correction in a negative (decreasing x_i) direction, and let m be the starting subscript for interpolation and correction in a positive (increasing x_i) direction. We must consider three cases when choosing j and m.

Case 1: $x_k = x_n$, then j = n and $m = \theta$ (not defined)

Case 2: $x_k = x_0$, then $j = \theta$ and m = 0

Case 3: x_k $k \neq 0,n$, then j = min(n,k+1) and m = max(0,k-1)

We perform the parabolic leapfrog corrections in a positive direction by choosing the set of points

$$\{(x_i,f_i),(x_{i+1},f_{i+1}),(x_{i+3},f_{i+3})\}$$

where m+3 \leq n and m \leq i \leq n-3, to find the point (x_{i+2},f_{i+2}) .

For leapfrogging in the negative direction we choose the set of points

$$\{(x_{i-3},f_{i-3}),(x_{i-1},f_{i-1}),(x_i,f_i)\}$$

where $j \ge 3$ and $j \ge i \ge 3$, to find the point (x_{i-2}, f_{i-2}) .

We note from the above that at least four points are required for this correction procedure and that the leapfrogging cannot always be done in both a positive and negative direction. Leapfrogging could not be used to correct a boundary value but can be used on interior points nearest to the boundaries. Because interpolation, by definition, requires known data to bound unknown points, correcting boundary values should not be required.

The Lagrange Interpolating Polynomial

A parabola was fitted through the set of "known" points in the leapfrogging stage using a procedure known as Lagrange interpolation. This technique was chosen because of its ease of use and its flexibility. A polynomial of degree n-1 is fitted through n points, so that if desired, one could include more points in the leapfrog's individual interpolations with minor modifications to the procedure.

Let

$$P(x) = \sum_{i=0}^{n-1} f_i L_i(x)$$

where P(x) = the interpolated value at x

 f_i = the value of the function to be approximated at known point x_i

 $L_i(x)$ = the ith Lagrange coefficient at x

 x_i = the ith known x value

n = the number of interpolating points

We define the Lagrange coefficient as

$$L_k(x) = \prod_{\substack{m=0 \ m \neq k}}^{n-1} \frac{x - x_m}{x_k - x_m}$$

for k = 0, 1, ..., n-1.

Henrici² gives the motivation for this technique and also a proof.

Cubic Splines

When a grid has been generated and corrected, cubic spline interpolation is used to supply data values at the desired unknown points. The cubic spline is a cubic polynomial that can be used with data sets of either fixed or variable mesh length. Cubic spline interpolation provides a minimum curvature fit to the data that is continuous through the second derivative and returns exact values at known points. For additional details see Sommers (1976).

APPLICATION

We first tested the grid generation scheme by reconstructing an existing grid of 51 rows and 51 columns with a data field consisting of a conical surface centered on gridpoint (26,26). The cone's surface formed a 30° angle with the plane of the grid and the data values ranged from 0.0 at (26,26) to slightly over 20.3 at the grid's corner points. The "known" points for the grid interpolation were chosen at random from every row, every other row, and every third row. The root-mean-square errors of the unknown points in each case are summarized in table 1. A FORTRAN IV program compiled by the FORTRAN H optimizing compiler and executed on an IBM 360 model 503 was used to perform the interpolation from the known points to the desired 51 X 51 grid. Execution times for each run are included in table 1 to give a feel for percentage change when the number of points is increased or decreased. The ratio of execution times would be fairly constant for most machines.

Table 1-Error and execution time for interpolation on a conical surface to a 51×51 grid

Number of known points	Root-mean- square error ¹	Execution time (min)
51	1.5179	1.53
26	2.5315	0.53
17	3.1588	0.46

¹ Root-mean-square errors were determined by taking the values at known points, dropping them out, interpolating to them as if they were unknown, and comparing the results.

As a second test, we selected a set of 71 standard reporting weather stations in the Southeastern United States, held out a subset as the "unknown" points, and interpolated the "known" points to the unknown points by generating a grid, using the same procedure as in the first test case, and then interpolating the gridded data to the unknown points. Figure 2 shows

the location of the 50 known points and the 21 unknown points. We performed the interpolation procedure on three variables: temperature, sea level temperature, and dewpoint.

During the initial attempt at interpolating the realworld data we discovered that the "shooting" error which sometimes arises when one is fitting curves became quite large and unacceptable. Shooting errors arise when the gradient between two known points is so large that the interpolating function trajectory continues to increase or decrease beyond reasonable limits at an unknown point downstream. The shooting error was corrected two different ways for comparison. The first and most obvious way to correct it was to limit the interpolated value to either the minimum or the maximum of the points being used for interpolation according to whether we undershot or overshot. The second was to fit the point using linear interpolation if shooting occurred. We ran the program both ways for six hourly observations between 1800 and 2300 GMT on April 16, 1976, and computed root-mean-square errors at each of the "unknown" points. Linear interpolation resulted in minimum root-mean-square errors in all cases analyzed. The results are given in table 2.

Table 2-Root-mean-square temperature, sea level temperature, and dewpoint errors at interpolated stations for hour observations between 1800 and 2300 GMT on April 16, 1976, using two methods for correcting shooting error

Correction method	Root-mean-square error for interpolated		
	Temperature	Sea level temperature	Dewpoint
		- °c	
Minimum/maximum limiting Linear	2.46	2.51	3.76
interpolation	2.42	2.43	3.38

The points used as the "known" points were chosen at random and the grid used for the interpolation was arbitrarily given 25 rows and columns. This grid size was chosen to hold computer core and cpu requirements down in an effort to speed up the debugging process by increasing the total number of runs per day. A careful analysis of the known data available and a judicious choice of the grid's mesh length based on the user's particular data base should

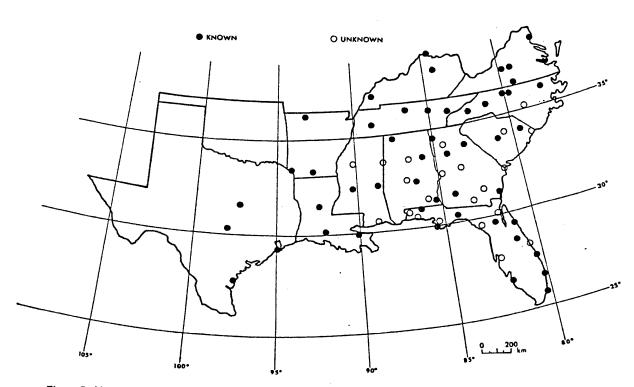


Figure 2—Meteorological data from "known" sites in the 13 southeastern States were interpolated to "unknown" sites in the six most southeasterly States.

yield even better results. These two considerations are the most important when using this particular scheme for interpolation of data.

FORTRAN IV routines which will perform the interpolation of randomly spaced data to grid, and which will interpolate gridded data to randomly spaced points, are available on request to: Director, Pacific Southwest Forest and Range Experiment Station, P.O. Box 245, Berkeley, California 94701, Attention: Computer Services Librarian.

NOTES

¹Sommers, W. T. 1976. Data interpolation by cubic splines. USDA Forest Serv. Res. Note PSW-313, 5 p. Pacific Southwest Forest and Range Exp. Stn., Berkeley, Calif.

² Henrici, P. 1964. *Elements of numerical analysis*. p. 183-193. John Wiley and Sons, New York.

³ Trade names and commercial enterprises or products are mentioned solely for information. No endorsement by the U.S. Department of Agriculture is implied.

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Attachment Q COMPARING METHODS OF FILLING IN MISSING DATA

Suppose one has a number of years of monthly average values of pollutant readings, and that in some one year, some of the monthly averages are missing. Number the months by $i=1,2,\ldots,12$. We will address our remarks to three different methods of filling in the missing values.

The first assumes that the seasonal pattern is fairly constant. Thus, if X_{ij} is the i^{th} monthly average

$$X_{ij} = P_i T_j(1+E_{ij})$$

where T_j is the total for the jth year, $\sum_i P_i = 1$, and as a consequence

$$T_{j} = \sum_{i} X_{ij} = T_{j} \sum_{i} P_{i}(1+E_{ij})$$

implies that the random errors E_{ij} satisfy

(1)
$$\sum_{i} P_{i} E_{ij} = 0$$
, all j.

We will assume that the ${\rm E}_{ij}$ are normally distributed with equal variances σ^2 and mean zero conditioned on (1).

Suppose that we have J years of good data, and in year j = o, we have good data in the months $I = \{i_1, \ldots, i_m\}$ and data missing in the other months whose set of indices we denote by I^C . A natural method of filling

-2-

in the missing data is to estimate the P_{i} from the J good years by taking

$$P_{i}^{A} = \frac{1}{J} \sum_{j>0} \chi_{ij} / T_{j}$$

$$= \frac{1}{J} \sum_{j>0} P_{i} (1 + E_{ij}) = P_{i} \left(1 + \frac{1}{J} \sum_{j>0} E_{ij}\right).$$

Assume also that the E_{ij} are independent for different j and independent for i before the conditioning (1).

Now estimate the total T_0 for the year with missing data by

$$\begin{split} \widehat{T_o} &= \frac{1}{I} \sum_{i \in I} X_{ij} / \widehat{P_i} \\ &= \frac{1}{I} \sum_{i \in I} P_i T_o (1 + E_{io}) / \widehat{P_i} \\ &= T_o \left(\frac{1}{I} \sum_{i \in I} (1 + E_{i,o}) P_i / \widehat{P_i} \right) . \end{split}$$

Now fill in the missing data by the estimate

$$\hat{X}_{io} = \hat{P}_{i} \hat{T}_{o}$$

We want to estimate the size of the overall bias over the missing months. This bias is defined as

$$B = \sum_{i \in I} c (\hat{X}_{io} - P_i T_o).$$

As our measure of bias, we will compute EB^2 . Now

$$B = \sum_{i \in I} c^{\hat{X}}_{io} - T_o \sum_{i \in I} c^{p}_{i}$$

Now

$$\hat{x}_{io} = P_i T_o \left(1 + \frac{1}{J} \sum_{j \geq 0} E_{ij} \right) \left(\frac{1}{I} \sum_{i \in I} (1 + E_{i,o}) P_i / \hat{P}_i \right)$$

Write

(2)
$$\frac{1}{I} \sum_{i \in I} (1 + E_{i0}) P_i / \hat{P}_i = 1 + \frac{1}{I} \sum_{i \in I} \frac{P_i - \hat{P}_i}{P_i^2} + \frac{1}{I} \sum_{i \in I} E_{i0} P_i / \hat{P}_i ,$$

and approximate

$$\frac{P_{i}-P_{i}^{\prime}}{P_{i}^{\prime}} \simeq \frac{P_{i}-P_{i}^{\prime}}{P_{i}} = -\frac{1}{J} \sum_{j>0} E_{ij}$$

$$\sum_{i \in I} E_{io} P_i / P_i \simeq \sum_{i \in I} E_{io}$$

to get the expression in (2) to be

$$1 - \frac{1}{JI} \sum_{\substack{j>0 \ i \in I}} E_{ij} + \frac{1}{I} \sum_{i \in I} E_{i0}$$

Therefore, to first-order terms

$$\hat{x}_{i,o} = P_i T_o \left(1 + \frac{1}{J} \sum_{j>0} E_{i,j} - \frac{1}{IJ} \sum_{j>0} E_{ij} + \frac{1}{I} \sum_{i \in I} E_{io} \right)$$

SO

$$x_{i,0}^{\uparrow} - P_{i} T_{o} = P_{i} T_{o} \left(\frac{1}{J} \sum_{j>0} E_{i,j} - \frac{1}{IJ} \sum_{j>0} E_{ij} + \frac{1}{I} \sum_{i \in I} E_{io} \right).$$

Summing over $i \in I^{\mathbb{C}}$ gives

$$B/T_0 \simeq \frac{1}{J} \sum_{\substack{j > 0 \\ i \in I}} P_i \ E_{i,j} - \frac{P}{I} \sum_{\substack{j > 0 \\ i \in I}} E_{ij} + \frac{P}{I} \sum_{i \in I} E_{i0}$$

where $P = \sum_{i}^{C} P_{i}$. Since $\sum_{i}^{C} P_{i} E_{ij} = 0$, we can write

$$\sum_{\substack{j>0\\i\in I^c}} P_i E_{ij} = -\sum_{\substack{j>0\\i\in I}} P_i E_{ij}$$

and rewrite

$$B/PT_{0} \simeq -\frac{1}{J} \sum_{\substack{j>0 \\ i \in I}} E_{ij} (\frac{Pi}{P} + \frac{1}{I}) + \frac{1}{I} \sum_{i \in I} E_{i0}.$$

Note that the two terms are independent, and have mean zero.

We can go a little further using a standard probability result, i.e., let Z_1, \ldots, Z_n be independent $N(o, \sigma^2)$ variables. Then

(3)
$$\mathbb{E}\left[\left(\sum_{j} \mathcal{L}_{j} \ Z_{j}\right)^{2} \left| \sum_{j} \beta_{j} \ Z_{j} = 0\right] = \sigma^{2}\left[\sum_{j} \mathcal{L}_{j}^{2} - \frac{\left(\sum_{j} \mathcal{L}_{j}^{\beta}\right)^{2}}{\sum_{\beta_{j}^{2}}}\right] \right] .$$

From the above, we have that

$$\frac{1}{P^{2}T_{0}^{2}} \quad EB^{2} = \frac{1}{J^{2}} E \left[\sum_{\substack{j>0\\i \in I}} E_{ij} \left(\frac{P_{i}}{P} + \frac{1}{I} \right)^{2} \middle| \sum_{i} P_{i} E_{ij} = 0 \right]$$

$$+ \frac{1}{I^{2}} E \left[\left(\sum_{i \in I} E_{io} \right)^{2} \middle| \sum_{i} P_{i} E_{io} = 0 \right].$$

Since the E_{ij} are independent for different j, the first term is simply

$$\frac{1}{J} E \left\{ \left[\sum_{i \in I} E_{ii} \left(\frac{P_i}{P} + \frac{1}{I} \right) \right]^2 \middle| \sum_{i} P_i E_{ii} = 0 \right\}.$$

By (3) above this equals

(5)
$$\frac{\sigma^2}{J} \left\{ \sum_{i \in I} \left(\frac{P_i}{P} + \frac{1}{I} \right)^2 - \frac{\left[\sum_{i \in I} P_i \left(\frac{P_i}{P} + \frac{1}{I} \right) \right]^2}{\sum_{i} P_i^2} \right\} ...$$

Using (3) again, the 2nd term in (4) is

(6)
$$\frac{1}{I^2} \sigma^2 \left[I - \frac{\left(\sum_{i} P_i\right)^2}{\sum_{i} P_i^2} \right] .$$

As a first approximation, take all $P_i = 1/12$. Then (5) becomes

$$\frac{\sigma^2}{J} \left(\frac{1}{I} + \frac{1}{I^c} \right)^2 \left[I - \frac{I^2}{12} \right]$$

$$= \frac{\sigma^2}{J} \cdot \frac{12}{II^c}.$$

Also (6) becomes

$$\frac{\sigma^2}{I^2} \cdot \frac{II^c}{12} = \sigma^2 \frac{I^c}{12I}$$

Noting that $P = I^{C}/12$ we get

$$EB^{2} = T_{0}^{2} \sigma^{2} \left[\frac{12}{JII^{c}} \cdot \frac{(I^{c})^{2}}{12^{2}} + \frac{I^{c}}{12I} \cdot \frac{I_{c}^{2}}{12^{2}} \right]$$
$$= \frac{\sigma^{2}T_{0}^{2}}{12} \left[\frac{I^{c}}{JI} + \frac{1}{12^{2}} \frac{I_{c}^{3}}{I} \right]$$

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(7)
$$= \frac{\sigma^2 T_0^2}{12} \left(\frac{I_c}{I} \right) \left[\frac{1}{J} + \left(\frac{I^c}{12} \right)^2 \right] .$$

On the other hand, suppose we fill in the missing values by averaging two adjacent years, i.e., take

$$\hat{x}_{i,0} = (x_{i,1} + x_{i,2})/2$$

$$= P_i [(T_1 + T_2)/2] + \frac{1}{2} P_i T_1 E_{i,1} + \frac{1}{2} P_i T_2 E_{i,2} .$$

Then the bias B is given by

$$B = \sum_{I^{C}} (X_{i,0} - P_{i} T_{0})$$

$$= \sum_{I^{C}} P_{i} \left(\frac{T_{1} + T_{2}}{2} - T_{0} \right) + \frac{1}{2} \sum_{i \in I^{C}} (T_{1} P_{i} E_{i,1} + P_{i} T_{2} E_{i,2}) .$$

Assuming again that $P_i = 1/12$, we get

$$EB^{2} = \left(\frac{I^{c}}{12}\right)^{2} E\left(\frac{T_{1}^{+}T_{2}}{2} - T_{0}\right)^{2} + \left(T_{1}^{2} + T_{2}^{2}\right) \cdot \frac{\sigma^{2}}{12^{2}} \frac{II^{c}}{12} \cdot \frac{1}{4} .$$

If we take $T_1 = T_2 = T_0$ to give a lower bound, then we have

(8)
$$EB^{2} \ge \frac{1}{2} T_{0}^{2} \sigma^{2} \frac{II^{c}}{12^{3}} .$$

A simple calculation shows that the expression in (7) is lower than the bound in (8) for $I^C \le 3$, for all $J \le 8$. In other words, for 3 or less months of missing data, if $J \le 8$, use the first method. If one assumes

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that $\left(\frac{T_1+T_2}{2}-T_0\right)^2$ is small, then for 4 or more months of missing data, use the 2_{nd} method.

If we only use one adjacent year to fill in, the expression for ${\rm EB}^2$ is

$$EB^{2} = \left(\frac{I^{c}}{12}\right)^{2} E(T_{1} - T_{0})^{2} + T_{1}^{2} \frac{\sigma^{2}}{12^{2}} \cdot \frac{II^{c}}{12} .$$

The lower bound with $T_1 = T_0$ is

(9)
$$EB^2 \ge T_0^2 \sigma^2 \frac{II^c}{12^3}$$
.

In this case, for J = 5, use the first method if 4 or less months are missing. For J = 7, it becomes 5 or less.